REM IV

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Black & Veatch ICF PRC Ecology and Environment ASSESSMENT OF THE TOXICITY OF COPPER, MERCURY, SELENIUM, SILVER AND THALLIUM IN SOIL AND PLANTS IN THE HELENA VALLEY OF MONTANA

for

EAST HELENA SITE (ASARCO) EAST HELENA, MONTANA

EPA Work Assignment No. 68-8L30.0

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TABLE OF CONTENTS

			Page
List	of Cor of Tab ary of		iii iv
1.0	1.1 1.2 1.3	DUCTION Purpose Scope Methods Site Description	1 1 1 2
2.0	BACKG SELEN	ROUND AND ELEVATED LEVELS OF COPPER, MERCURY, IUM, SILVER AND THALLIUM IN SOILS AND PLANTS.	4
	2.1	Copper Levels in Soils and Plants 2.1.1 Total copper levels in soils 2.1.2 Copper levels in plants	4 5 10
	2.2	Mercury Levels in Soils and Plants 2.2.1 Total mercury levels in soils 2.2.2 Mercury levels in vegetation	12 18 28
	2.3	Selenium Levels in Soils and Plants 2.3.1 Total selenium levels in soils 2.3.2 Selenium levels in plants	29 33 38
	2.4	Silver Levels in Soils and Plants 2.4.1 Total silver levels in soils 2.4.2 Silver levels in plants	39 43 43
	2.5	Thallium Levels in Soils and Plants 2.5.1 Total thallium levels in soils 2.5.2 Thallium levels in plants	49 51 52
3.0	HAZARI SILVEI	D LEVEL DEVELOPMENT FOR COPPER, MERCURY, SELENIUM R AND THALLIUM IN SOILS AND PLANTS	60
	3.1 3.2 3.3 3.4 3.5	Copper Hazard Levels Mercury Hazard Levels Selenium Hazard Levels Silver Hazard Levels Thallium Hazard Levels	60 66 67 69 70
4.0	REFERI	ENCES CITED	73

LIST OF TABLES

Numi	per ·	Page
1	Background total copper levels in soils	7
2	Elevated total copper levels in soils	9
3	Background copper levels in plants	13
4	Elevated copper levels in plants	16
5	Background total mercury levels in soils	22
6	Elevated total mercury levels in soils	27
7	Background mercuy levels in plants	30
8	Elevated mercury levels in plants	32
9	Background total selenium levels in soils	34
10	Elevated total selenium levels in soils	37
11	Background selenium levels in plants	40
12	Elevated selenium levels in plants	42
13	Background total silver levels in soils	44
14	Elevated total silver levels in in soils	45
15	Background silver levels in plants	46
16	Elevated silver levels in plants	50
17	Background total thallium levels in soils	53
18	Elevated total thallium levels in soils	54
19	Background thallium levels in plants	56
20	Elevated thallium levels in plants	57
21	Total concentrations of selected trace elements	
	considered phytotoxically excessive in soils	61
22	Plant tissue levels considered to be phytotoxic	62
23	Proposed hazard levels for soils and plants in the	
	Helena Valley study area	63

Units

kg kilogram; kg = 103 g
g gram = 10-3 kg
mg milligram; mg = 10-3 g
ug microgram; ug = 10-3 mg
ng nanogram; ng = 10-3 ug
L liter; L = 1 dm3
ml milliliter; ml = 10-3 L

Symbols

parts per million = ug/g = mg/kg ppm parts per billion = 10-3 ppm, ng/g = ug/kgppb ug/g microgram/gram mg/kg milligram/kilogram mg/L milligram/liter ug/L microgram/liter microgram/milliliter uq/ml ng/ml nanogram/milliliter

Acronyms

AAS Atomic absorption spectrophotometry AOAC Association of Official Agricultural Chemists AWT Ash weight basis CCM Copper carbonate method CEC Cation exchange capacity DTPA Diethylenetriaminepentaacetic acid DW Dry weight basis EDTA Ethylenediaminetetraacetic acid EPA Environmental Protection Agency EPA CV Environmental Protection Agency cold vapor method ES Emission spectrographic FLAAS Flameless atomic absorption spectrophotometry GLC Gas liquid chromatography INAA Instrumental neutron activation analysis IPAA Instrumental photon activation analysis MMC Methyl mercuric chloride MMH Methyl mercuric hydroxide MYC Mycorrhiza ND Not determined NOAA National Oceanic and Atmospheric Administration NR Not reported NRC National Research Council OM Organic matter content PH Negative logarithm, base 10, of H+ concentration PMA Phenyl mercuric acetate RNAA Radiochemical neutron activation analysis SSMS Spark source mass spectrometry USDA United States Department of Agriculture USGS United States Geological Survey WW Wet weight basis XRFL X-ray fluorescence Yield reduction YR

This document consists of a literature review and presents candidate levels of copper, mercury, selenium, silver and thallium for assessment of selected environmental hazards associated with the East Helena smelter located in the Helena Valley of Montana. This document is the second of two volumes. Volume one contains similar hazard levels for arsenic, cadmium, lead and zinc in addition to an evaluation of the hazard to livestock from these four elements. Candidate hazard levels presented in this report have been developed specifically for the East Helena, Montana smelter site. The use of this document for evaluation of other sites should be attempted only with proper consideration of site specific conditions.

1.1 Purpose

This document is a literature review from which proposed hazard levels have been developed to assess risk from chemical element levels found in soils and crops present in the vicinity of the East Helena smelter. These hazard levels will enable determination of the potential danger to these agricultural resources.

1.2 Scope

The scope of this document is confined to the metals copper, mercury, selenium, silver and thallium and their toxic effects and levels of accumulation in soils and plants. This document does not contain a review of relevant literature pertaining to extractable levels of these metals in soils.

1.3 Methods

Portions of the literature that are presented in this document were procured through the use of a computer search utilizing numerous data bases including AGRICOLA, BIOSIS, CAB Abstracts, CRIS-USDA, ENVIROLINE, MEDLINE, NTIS, Pollution Abstracts, SCISEARCH and Water Resources Abstracts. Conventional

library methods have also been employed for researching abstracts, periodicals and other materials. The authors are cognizant of the limitations of solution culture and greenhouse studies but for some aspects of the five metals reviewed, these are the only data available.

Background values presented were taken directly from the scientific literature. Phytotoxic levels were chosen through; 1) a review of levels reported to be phytotoxic in experimental studies and 2) a comparison of the reported experimental results with phytotoxic levels established by others. The scarcity of data precluded establishment of an upper tolerable concentration for some of these elements. Scientific literature that most closely approximated conditions present in the Helena Valley were emphasized more for hazard level selection. For example, a toxic soil level for wheat on calcareous loamy soil was considered more applicable than a toxic soil level for cabbage on a sandy acid soil. Once hazard levels were developed they were compared to means and ranges of soil/plant elemental levels measured in the Helena Valley and control sites.

All values reported in this document are on a dry weight basis unless otherwise indicated.

1.4 Site Description

The Helena Valley is located in west central Montana and trends in a west northwest direction. It is 35.4 km (22.1 mi) long and 17.1 km (10.7 mi) wide. The valley is bounded on the northeast by the Big Belt Mountains, on the south by the Elkhorn Mountains and the Boulder Batholith, and on the west by mountains forming the continental divide. Lower portions of the valley are occupied by Lake Helena and Hauser Lake formed by Hauser dam on the Missouri River. Elevations range from 1,113 m (3650 ft) mean sea level at Hauser Lake to 2,560 m (8,400 ft) in the surrounding mountains. Geological materials on the valley floor consist of quaternary and tertiary sediments that are consolidated to poorly consolidated. Soils are moderately calcareous and composed of silt and clay (Miesch and Huffman 1969). Soil profiles are

poorly to moderately developed on both quaternary and tertiary parent materials. The Helena Valley is semi-arid and receives from less than 25.4 cm (10 in) to less than 36 cm (14 in) of annual precipitation. The adjacent mountains receive up to 76.2 cm (30 in) of annual precipitation (U.S. Soil Conservation Service 1981). The climate is modified continental with an average annual temperature of 6.3°C (43.3°F) (National Oceanic and Atmospheric Administration (NOAA 1983). Average January and July temperatures at Helena are -8°C (18.1°F) and 20°C (67.9°F) respectively (NOAA 1983). Agricultural crops in the Valley are alfalfa, small grains (usually wheat, barley and some oats) and range land.

The Helena Valley is the site for two incorporated cities: Helena and East Helena with approximate populations of 23,900 and 2,400 respectively (1980 census). The two cities are located 6.4 (4 mi) and 1 km (0.6 mi) from the smelter complex, respectively.

The valley has been the site of a lead smelter since the Helena and Livingston facility was built in East Helena in 1888. The smelter was purchased by its present owner (American Smelting and Refining Company) in 1899. The Anaconda Company built a zinc plant adjacent to the smelter in 1927 to recover zinc from waste products. In 1955 the American Chemet Company constructed a paint pigment plant utilizing zinc oxide from the zinc facility.

2.0 BACKGROUND AND ELEVATED LEVELS OF COPPER, MERCURY, SELENIUM, SILVER AND THALLIUM IN SOILS AND PLANTS.

Varing amounts of research data are available for copper, mercury, selenium, silver and thallium. For copper and, to a lesser extent, mercury a large volume of work has been completed in reference to sewage sludge disposal problems. Our understanding of selenium has benefited from the studies of selenium accumulator plants and their adverse effects on livestock. Little data are available to accurately evaluate levels of silver and thallium found in soils and plants. Copper is the only one of these elements considered essential for higher plants (Kabata-Pendias and Pendias, 1984). Sections 2.1 through 2.5 discuss unique characteristics of each metal reviewed and levels found in soils and plants.

2.1 Copper Levels in Soils and Plants

Copper is one of the most studied heavy metals. Extensive literature has been published concerning the role of copper in animals and plant nutrition, sewage sludge disposal, and environmental pollution from industrial sources. Study of the beneficial and toxilogical effects of copper in agricultural crops date from research published by Grossenbacher (1916), Floyd (1917) and Forbes (1917).

The total concentration of copper in the earth crust has been reported at approximately 50 ppm (National Research Council, NRC 1977). Bowen (1966) reported copper levels in igneous rock, shale, sandstone and limestone as 55 ppm, 45 ppm, 5 ppm and 4 ppm respectively. The copper content of shale, bituminous shale, sandstone and limestone and dolomite were reported by others as 35 ppm, 70 ppm, 30 ppm and 6 ppm respectively (Wedepohl and Zemann 1974). These authors also reported a copper concentration in coal as 17 ppm. Copper is most abundant in mafic and intermediate rocks and minerals such as biotite and pyroxene (Cox 1979, Mitchell 1971, Thornton 1979). It is usually found as simple and/or complex sulfides, many of which are easily soluble,

especially in acid environments (Kabata-Pendias and Pendias 1984, NRC 1977). Copper also occurs as a native metal (Cox 1979).

Haque and Subramanian (1982) reported atmospheric emissions of copper as 18,500 and 56,000 metric tonnes per year for natural and anthropogenic sources respectively. It was estimated that 65 percent of the natural emissions occur from windblown dusts and that "vegetative exudates account for the bulk of the remainder" (Nriagu 1979). Metallurgical processing and wood combustion have been reported as the major sources of anthropogenic copper at levels of 19,800 and 11,500 tonnes per year, respectively. "About 95 percent of the anthropogenic copper emissions comes from point sources such as smelters, utility plants and incinerators" (Nriagu 1979).

2.1.1 Total copper levels in soils

A complete discussion of the role and function of copper in soils and plants is beyond the scope of this document. The following brief discussion is provided to help the interpretation of reported soil levels.

The copper content of most soil is determined in part, from copper present in parent material. The soil level is modified to varying degrees by pedogenetic processes (Thornton 1979). These processes include climatic factors which determine the amount of weathering and degree of soil formation, topography, soil pH, the redox potential and the organic matter content (Baker 1974).

The form of copper in soils remains somewhat obscure. Although copper occurs in two valence states, Cu^{+1} and Cu^{+2} , copper in soil is almost exclusively in the Cu^{+2} form (Thornton 1979).

The three soil parameters most likely to control copper availability to plants are pH, cation exchange capacity (CEC) and organic matter content (OM). The soil pH is the parameter most consistently identified in the literature as controlling metal availability to plants. All microelements, with the exception of molybdenum and selenium "are more labile at low pH due to hydrolysis of hydroxide species and (increased) solubility of

other solid phase minerals such as carbonates and phosphate (Logan and Chaney 1983). A pH level >6.5 is considered to be effective in reducing the plant availability of soil copper and other metals (Chaney 1973, CAST 1976). Copper is sorbed or bound more strongly to soil colloids and organic matter than are many other cations (Reuther and Labanauskas 1966, Thornton 1979). Leeper (1972) suggested that soil CEC be used as an index to determine the amount of metals that can be added to a soil without producing phytotoxicity. This index may be more applicable to smelter pollution than it is to sewage sludge due to the sorption properties of sludge which dominate the CEC and OM properties of the soil to which it is applied (Corey 1981). humic, fulvic and tannic acids of organic matter form stable compounds with copper and other metals (Stumm and Morgan 1970). Stevenson and Ardakani (1972) have reported that copper organometallic complexes are more stable than similar complexes of lead, iron, nickel, manganese, cobalt and zinc at a pH of 5. Nickel and copper are typically associated with soils high in organic matter content (Hazlett et al. 1983).

The background total soil copper concentration can range from 1 to 300 ppm with means generally in the range of 10 to 50 ppm (Table 1). Kubota (1983) reported a range and mean of 2-137 ppm and 30 ppm respectively for Western United States valley fill materials.

Elevated copper levels in soils are less well documented than are background data (Table 2). Much of these data related to elevated copper levels in soil have been associated with sewage sludge disposal problems. The interactions of other metals with copper in sludges and the effect of sludge organic matter make interpretation of these data difficult. Elevated copper data reported in reviewed literature ranged from typical background levels to the 2254 ppm copper found in abandoned open ore roasting areas (Hogan et al. 1977). Selection of hazard levels for elevated copper concentrations in soils is presented in Section 3.1.

Table 1. Packground total copper levels in soils.

									یے ')
		Level	Hazard	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference	
Medium	Use (pp	m DW) Means in ()	Response	Pathway	Receptor					
Minnesota Soils	Not given	16-50 (26)	Background	NR	NR		Tri-Acid AAS	Field	Pierce et al. (1982)	
Japanese Soil	Not given	4.4-176 (34)	Background	NR	NR	NR	Not given	Field	Kitagishi and Yamane (1981)	
Organic Muck Soils	Uncultivated	2-27	Background	Plant Uptake	NR	Maturity	AAS	Field	Czuba and Hutchinson (1980)	
Surface - Ontario	Truck Farm	72-213	Background	Plant Uptake	Vegetables	Maturity	AAS	Field	Czuba and Hutchinson (1980)	
O: ganic Muck Soils								E1.74		
40-48 cm - Ontario	Truck Farm	1-123	Background	Plant Uptake	Vegetables	Maturity	AAS	Field	Czuba and Hutchinson (1980)	
Ontario Soils	Crops	(15.9)	Background	NR	NR	NR	AAS	Field	Czuba and Hutchinson (1980)	
Piedmont Soils	Forage	13-191 (52)	Background	Plant Uptake	Legumes/grasses	Not given		Field	Price et al. (1955)	
California Soils	Crops	8-112 (54)	Background	Plant Uptake	Legumes/grasses	Not given				
							Colormetri-	m:-12	W. b = t = (1002)	
							cally AAS	Field	Kubota (1983)	
Western US						33-2		Field	Wharta (1002)	
Valley Fill	Crops	2-137 (30)	Background	Plant Uptake	Legumes/grasses	Not given		rieid	Kubota (1983)	
Glacial Drift				_9				Field	Kubota (1983)	
N. Central and	Crops	1-119 (17)	Background	Plant Uptake	Legumes/grasses	Not given		Field	Kubota (1983)	
New England	Crops	1-179 (24)	Background	Plant Uptake	Legumes/grasses	Not given		Field	Kubota (1983)	
Alluvium (Calif)	Crops	1-55 (15)	Background	Plant Uptake	Legumes/grasses	Not given		riera	Rubota (1983)	
Coastal Plain		T GOD BOOK				Not given		Field	Kubota (1983)	
(SE.US), NC, SC	Crops	8-34 (19)	Background	Plant Uptake	Legumes/grasses	Not given		rieid	Kubota (1903)	
Coastal Plain		a words workers	100 IN IN			Not given		Field	Kubota (1983)	
Fl, NC, SC	Crops	1-13 (5)	Background	Plant Uptake	Legumes/grasses	Not given		rieid	Rubota (1983)	
Western US Soils	Native Range				Native Range/	Not given	ND	Field	Shacklette and Boerngen (1984	Y
	Crops	2-300 (21)	Background	Plant Uptake	Crop Plants	Not given	MK	rield	Shacklette and Boerngen (1904	,
US Soils	Native Range			whose nearly	Native Range/	Not given	NR	Field	Shacklette and Boerngen (1984	1
	Crops	(17)	Background	Plant Uptake	Crop Plants	NA NA	AAS	Field	Connor et al. (1976)	,
Surface Soil/		14 (Geo. Mean)	Background	Plant Uptake	Sagebrush	NA	AAS	Field	Connor et al. (1976)	
Powder River Basin	Native Range		Background	Plant Uptake	Sagebrush Sagebrush	NA	AAS	Field	Connor et al. (1976)	
Subsoil/Powder		16 (Geo. Mean)	Background	Plant Uptake Plant Uptake	Sagebrush	NA NA	AAS	Field	Connor et al. (1976)	
River Basin	Native Range		Background	Plant Uptake	Agrostis gigante		AAS	Field	Hogan et al. (1977)	
	Native Plants		Background	Plant Optake	Agroscis gigance	a nacarrej			magen at any grown,	
Organic Soils		29.5-111.0	Background	Plant Uptake	Crops	NR	AAS	Field	Ishida and Suda (1976)	
0-15 cm	Crops	(65.0) 2.1-123.0 (20.2)		Plant Uptake	Crops	NR	AAS	Field	Ishida and Suda (1976)	
Sandy Soils 0-15 cm	Crops	3.8-144.0 (25.5)		Plant Uptake	Crops	NR	AAS	Field	Ishida and Suda (1976)	
Loam Soils 0-15 cm	Crops	9.5-77.2 (16.7)		Plant Uptake	Crops	NR	AAS	Field	Ishida and Suda (1976)	
Clay Soils 0-15 cm	Crops	2.1-144.0 (21.5)		Plant Uptake	Crops	NR	AAS	Field	Ishida and Suda (1976)	
All Ontario Soils	Crops	2.1-144.0 (21.3)	Background	riant opeane	01000				The state of the s	
Canadian Shield	Uncultivated	(11)	Background	Plant Uptake	NR	NR	AAS	Field	McKeague and Wolynetz (1980)	
Soils		(17)	Background	Plant Uptake	NR	NR	AAS	Field	McKeague and Wolynetz (1980)	
Canadian Appalachian St. Lawrence Lowland		(19)	Background	Plant Uptake	NR	NR	AAS	Field	McKeague and Wolynetz (1980)	
Canadian Interior	15	(1))	Buckground	. ranc speak						
Plains		(21)	Background	Plant Uptake	NR	NR	AAS	Field	McKeague and Wolynetz (1980)	
Canadian Cordilleran		(46)	Background	Plant Uptake	NR	NR	AAS	Field	McKeague and Wolynetz (1980)	
16 Manitoba Soils	Agriculture	(25) A Hor.	Background	Plant Uptake	NR	NR	AAS	Field	Mills and Zwarich (1975)	
16 Manitoba Soils	Agriculture	(23) C Hor.	Background	Plant Uptake	NR	NR	AAS	Field	Mills and Zwarich (1975)	
Michigan - Sand	Woodland	2.8	Background	Plant Uptake	NR	NR	AAS	Field	Klein and Russell (1973)	
Residential Soils	Lawns and	2.0	Buckground							
	Woodlands	(8.0)	Background	NR	NR	NR	AAS	Field	Klein (1972)	
Grand Rapids, MI Agricultural Soils,	Woodtailus	(0.0)	2232.00							
Michigan	Crops	(8.8)	Background	NR	NR	NR	AAS	Field	Klein (1972)	
Industrial Soils	Industrial	10.01	Duc.igr ou.i.s							
Grand Rapids, MI	Sites	(16.3)	Background	NR	NR	NR	AAS	Field	Klein (1972)	
Airport Soils	31163	(10.4)	Background	NR	NR	NR	AAS	Field	Klein (1972)	
Cottenham Sandy Loam	Onions	3.9 W/MYCA	Background	Plant Uptake	Onions	5 weeks	AAS	Greenhouse	Gildon and Tinker (1983)	
Cottenham Sandy Loam		2.8 WO/MYCA	Background	Plant Uptake	Onions	5 weeks	AAS	Greenhouse	Gildon and Tinker (1983)	
Correntiam Sandy Doan	. Onions	2.0 40/1110								

Table 1. Background total copper levels in soils, continued.

		Level	Hazard	Exposure				Study	
Medium	Use (ppm DW) Means in ()	Response	Pathway	Receptor	Duration	Method	Setting	Reference
Piedmont Weathered				name or aden is nowing		NR	AOAC	Field	Price et al. (1955)
Bedrock	NR	13-191 (52)	Background	Plant Uptake	Forage	NR NR	NR	Field	Bowen (1966)
Worldwide	NR	2-100 (20)	Background	NR	NR	N R	AAS	Field	Cataldo and Wildung (1978)
Ritzville Silt Loam	Agricultur	re 31	Background	Plant Uptake	Crops	NR	SSMS	Field	Ure and Bacon (1978)
Aberdeenshire, UK	NR	10-21	Background	NR	NR	WK.	00		
Yakima Co. WA	100,000			07700 M 107 107000 1000	European	Maturity	ES	Field	Shacklette (1980)
pH (7.9)	Grapes	20-30 (22)	Background	Plant Uptake	Grapes	Macuricy	БО		
San Jaquin Co. CA	A STATE OF S			Personal Committee Committ	European	Maturity	ES	Field	Shacklette (1980)
pH (6.4)	Grapes	15-50 (27)	Background	Plant Uptake	Grapes	Macurity	20		
Berrien Co. MI						Maturity	ES	Field	Shacklette (1980)
рн (6.6)	Orchards	10-30 (18)	Background	Plant Uptake	Apples	Macuitcy	,		
Wayne Co. NY			180		× ×	Maturity	ES	Field	Shacklette (1980)
pH (5.5)	Orchards	15-30 (19)	Background	Plant Uptake	Apples	Macuiley	20		
Gloucester Co. NJ						Maturity	ES	Field	Shacklette (1980)
pH (5.5)	Orchards	15-30 (20)	Background	Plant Uptake	Apples	Macuilly	20		
Yakima, Co. WA				0000 to 120 to 14000	******	Maturity	ES	Field	Shacklette (1980)
рн (6.6)	Orchards	20-70 (36)	Background	Plant Uptake	Apples	Macuilly	ьь		
Mesa Co. CA				**************************************		Maturity	ES	Field	Shacklette (1980)
pH (7.8)	Orchards	20-50 (31)	Background	Plant Uptake	Apples	Macurity			
San Joaquin Co. CA					4	Maturity	ES	Field	Shacklette (1980)
рн (6.8)	Orchards	15-150 (100)	Background	Plant Uptake	Peaches	Macuity	20		
Mesa Co. CO	01011010					Maturity	ES	Field	Shacklette (1980)
рн (7.7)	Orchards	20-30 (24)	Background	Plant Uptake	Peaches	Haculicy	20		A Annual Control of the Control of t
Mesa Co. CO	0200000	00000 00000 00 00000 00		207 OF 187 NO .	N	Maturity	ES	Field	Shacklette (1980)
pH (8.0)	Orchards	15-100 (28)	Background	Plant Uptake	Pears	Macuilly	55		
San Joaquine Co. CA	020110200			orner to successionation		Maturity	ES	Field	Shacklette (1980)
pH (7.0)	Orchards	150-300 (240)	Background	Plant Uptake	Pears	Macurity	55		
Yakima Co. WA	0201142	140				Maturity	ES	Field	Shacklette (1980)
pH (6.3)	Orchards	30-70 (44)	Background	Plant Uptake	Pears	Macuilty	20		
Wayne Co. NY	220110200				21.5	Maturity	ES	Field	Shacklette (1980)
pH (6.6)	Orchards	7-20 (13)	Background	Plant Uptake	Pears	maturity	20		
Berrien Co. MI	020.10200	· mas Acting	000 10 200		12000000000000000000000000000000000000	Maturity	ES	Field	Shacklette (1980)
pH (5.4)	Orchards	15-50 (25)	Background	Plant Uptake	Pears	maturity			

A Mycorrhiza

Table 2. Elevated total copper levels in soils.

		Level	Hazard	Exposure	-			Study	Deference
Medium	Use (ppm DW)	Response	Pathway	Receptor	Duration	Method	Setting	Reference
re Roasting Bed	Agrostis								
		220-2254 1	& Ground Veg. Cover	Plant Uptake	NR	Maturity	AAS	Field	Hogan et al. (1977)
lo Loam	Bush Bean	500	83 % YR	CuSO ₄	Leaf	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
lo Loam	Bush Bean	500	69 % YR	CuSO	Stem	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
rsaw Sandy Loam	Corn	343	60 % YRA	CuCl2/Sludge	Above Ground				
					Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
rsaw Sandy Loam	Rye	343	39 % YRA	CuCl ₂ /Sludge	Above Ground				
	70 4 cm				Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
rsaw Sandy Loam	Corn	343	59 % YRA	CuCl ₂ /Sludge	Above Ground				
tour buildy in the		- 1.5		The the account of the country of th	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
lo Loam	Bush Bean	200	26 % YR	CuSO ₄	Leaf	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
lo Loam	Bush Bean	200	14 % YR (N.S.)	Cuso4	Stem	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
rsaw Sandy Loam	Corn	194	41 % YRA	CuCl2/Sludge	Above Ground				
roam camer moun					Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
rsaw Sandy Loam	Rye	194	29 % YRA	CuCl ₂ /Sludge	Above Ground				
risaw bandy soum		***			Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
arsaw Sandy Loam	Corn	194	51 % YRA	CuCl ₂ /Sludge	Above Ground				ALTERNATION PRODUCTION OF THE PROPERTY OF THE
itsaw bandy boam	COLII			220-2, 22-2-5	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
rsaw Sandy Loam	Corn	150	68 % YRB	CuCl ₂	Above Ground				
itsaw Bandy Loam	COLII	130	00 0 1	cuciz	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975b)
rsaw Sandy Loam	Rye	150	43 % YRB	CuCl ₂	Above Ground		200		
Isaw Sandy Loam	Nyc	130	15 0 11.	04012	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975b)
rsaw Sandy Loam	Corn	150	61 % YRB	CuCl ₂	Above Ground				Section (Section 2)
Isaw Sandy Loam	COLII	130	OI O IN	cuciz	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975b)
rsaw Sandy Loam	Corn	120	45 % YRA	CuCl ₂ /Sludge	Above Ground		100		
Isaw Salidy Loam	COLII	110	15 0 11	00012/010090	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
rsaw Sandy Loam	Rye	120 1	4 % Yield IncreaseA	CuCl ₂ /Sludge	Above Ground				
irsaw Salidy Loam	Ky C		T TICLE INCICAGE	00012/010090	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
arsaw Sandy Loam	Corn	120	44 % YRA	CuCL ₂ /Sludge	Above Ground				
irsaw sandy Loam	COLII	120	44 0 IN	cuchy, brudge	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a)
olo Loam	Bush Bean	100	12 % YR (N.S.)	Cu SO 4	Leaf	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
lo Loam	Bush Bean	100	0.8 % YR (N.S.)	CUSO	Stem	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
Sandy Soil"	Per Rye Grass		50 % YR	Cu Salts	Shoot	4 weeks	AAS	Greenhouse	Dijkshoorn et al (1979)
	Onions	75 W/MYCC	3.7 % YR	CuSO4 3H20	Leaves	5 weeks	AAS	Greenhouse	Gildon and Tinker (1983)
ttenham Sandy Loam	Onions	75 WO/MYCC	11.5 % YR	CuSO4 3H20	Leaves	5 weeks	AAS	Greenhouse	Gildon and Tinker (1983)
ottenham Sandy Loam	Plantain	56	50 % YR	Cu Salts	Shoot	6 weeks	AAS	Greenhouse	Dijkshoorn et al (1979)
Sandy Soil"	White Clover	52	50 % YR	Cu Salts	Shoot	8 weeks	AAS	Greenhouse	Dijkshoorn et al (1979)
Sandy Soil"		50	17 % YR (N.S.)	CuSO ₄	Leaf	17 days	ES		Wallace et al. (1977a)
lo Loam	Bush Bean	50 .	1.1 % YR (N.S.)	CuSO4	Stem	17 days	ES		Wallace et al. (1977a)
olo Loam	Bush Bean		8 % Yield Increase	Cu304	Above Ground	I' days	55	Greenmouse, sorr rose	
arsaw Sandy Loam	Corn	40 0	(crop 1)	Sludge	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975b)
lan annance	-	42 .		studge	Above Ground	O WEEKS	mio	oreeooo	
rsaw Sandy Loam	Rye	46 9	6 % Yield Increase	n13	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975b)
			(crop 2)	Sludge	Above Ground	o weeks	MAD	Greennouse	cumingham ec al. (1575b)
rsaw Sandy Loam	Corn	46	17 % YR	Cludge		6 wooks	AAS	Greenhouse	Cunningham et al. (1975b)
	The second second		(crop 3)	Sludge	Biomass	6 weeks		Greenhouse	Gildon and Tinker (1983)
ottenham Sandy Loam	Onions	30 W/MYCC	2.8 % YR	Cuso4 3H20	Leaves	5 weeks	AAS	Greenhouse	Gildon and Tinker (1983)
ttenham Sandy Loam		30 WO/MYCC	16 % YR	CuSO4 3H20	Leaves	5 weeks	AAS	Greenhouse	Gildon and Tinker (1983)
ottenham Sandy Loam		15 W/MYCC	5.5 % YR	CuSO4 3H20	Leaves	5 weeks	AAS	Greenhouse	Gildon and Tinker (1983)
ottenham Sandy Loam	Onions	15 WO/MYCC	5.7 % YR	CuSO4 3H20	Leaves	5 weeks	AAS	Greenhouse	Gildon and Tinker (1983)
ottenham Sandy Loam	Onions	5 W/MYCC	7 % Yield Increase		Leaves	5 weeks	AAS		Gildon and Tinker (1983)
ottenham Sandy Loam	Onions	5 WO/MYCC	3.8 % YR	CuSO ₄ 3H ₂ O	Leaves	5 weeks	AAS	Greenhouse	GIIGOII GIIG IIINEL (1903)

A Other metal levels: Zn - 410 ppm, Cr - 404 ppm, Ni - 37 ppm B Other metal levels: Zn - 300 ppm, Cr - 350 ppm, Ni - 15 ppm C Mycorrhiza

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2.1.2 Copper levels in plants

Copper is known to be an essential nutrient for both plants and animals and, except for molybdenum, is the least abundant essential nutrient in soil. Most problems involve copper deficiency in plants or animals and copper toxicosis is uncommon except in mineralized areas or zones polluted by mining and smelting activities (Gough et al. 1979, Hutchinson 1979).

Uptake of copper increases with increased copper levels in soil (Wallace et al. 1977a). Absorption of copper is thought to be active but passive absorption may also occur, especially under conditions of high soil copper concentrations (Kabata-Pendias and Pendias 1984).

Copper concentrations have been found to be markedly higher in plant roots as opposed to above ground parts (Agarwala et al. 1977, Chino 1981, Forbes 1917 and Jarvis 1978). Plant roots exhibit a strong capability to hold copper under both deficiency and toxic conditions (Kabata - Pendias and Pendias 1984). Jarvis (1978) reported that up to 96 percent of the total plant copper content of ryegrass is retained by roots under high uptake conditions and the copper held by roots is not available to ryegrass shoots even after a further supply of copper is withdrawn. Toxic concentrations in root are only negligibly translocated to the above ground biomass (Bennett 1971) probably because copper in plant roots is insoluble in its association with cell walls (Jarvis 1978).

Copper toxicity in plants is produced from several factors:

1) root tissue damage, which restricts root extension, membrane permeability and inhibits translocation of iron (Bennett 1971, Kabata - Pendias and Pendias 1984); 2) Peroxidation of chloroplast membrane lipids and inhibition of photosynthetic electron transport (Kabata - Pendias and Pendias 1984); and 3) immobilization of copper in cell walls, cell vacuoles and nondiffusible copper-protein complexes (Kabata - Pendias and Pendias 1984). Elevated copper concentrations also adversely affect potassium uptake in cereal grains (Bujtas and Cseh 1981).

The first symptom of copper toxicity is depressed growth (Dijkshoorn et al. 1979) and retarded germination, seedling growth, and root development (Forbes 1917, Chapman et al. 1940, Reuther et al. 1952, Reitz and Shimp 1953, Dekock 1956). Copper toxicity results in an induced iron chlorosis, depressed tillering and thick, short, barbed-wire roots (Agarwala et al. 1977, Bennett 1971, Chino 1981, Reilly and Reilly 1973). Sensitive crops are cereals, legumes, spinach and citrus seedlings.

Copper has been shown to be synergistic with zinc, nickel and cadmium in solustion culture experiments using bush beans (Wallace and Romney 1977c). "Copper, nickel and cadmium were more toxic together than any one alone". These authors also noted a synergistic effect (decreased levels of phosphorus, zinc and iron in bush bean roots) when copper and nickel were applied together in solution culture experiments.

A major factor influencing copper toxicity in plants is the variation exhibited by different plant species in uptake and susceptibility to copper toxicosis. Leguminous plants seem particularly sensitive to high concentrations of copper. This is due to the inhibitory effect of copper on root nodulation and fixation (Vesper and Weidensaul 1978, Porter and Sheridan 1981). Vesper and Weidensaul (1978) reported that copper at all levels tested had adversely impacted dry weights of stems and foliage. Copper treatments of 5 and 10 ppm decreased nitrogen fixation 39 and 46 percent respectively. All copper levels reduce nodulation. Porter and Sherdian (1981) reported nitrogen fixation was eliminated in alfalfa at solution concentrations of 100 ug copper/ml.

In contrast there are some plants which are tolerant to elevated copper levels. Wallace et al. (1977e) found that 51.2 ppm copper (in shoots) had no adverse impact upon the vegetative yield of rice plants. Hogan and Rauser (1979) found that a 50 percent reduction in yield occurred in non-tolerant clones of Agrostics gigantea at concentrations of 8 mmol/m-3, while this level of reduced growth was not reached by the tolerant clone until concentrations exceeded 40 mmol/m-3. Haque and Subramanian

(1982) reported that the yield of perennial ryegrass was reduced after the dry matter accumulation of copper exceeded 40 ppm.

Typical background concentrations of grasses and legumes are 5 and 15 ppm respectively (Table 3). Price et al. (1955) measured a copper concentration range of 1.5 ppm (timothy) to 29.0 ppm (red clover) in the Piedmont area of Virginia. Elevated levels of copper in vegetation range up to 1457 ppm found in the roots of copper tolerant clones of Agrostis gigantea (Hogan and Rause, 1979). These authors reported shoot copper concentrations of 487 to 801 ppm in the tolerant clones of this species. Levels considerably below these concentrations are phytotoxic to many plants (Table 4). Selection of phytotoxic criteria for copper in plants is discussed in section 3.1.

2.2 Mercury Levels in Soils and Plants

Mercury, the only liquid metal at normal temperatures of the earth's surface, is present in trace amounts in most geological materials, soils and plants (Connor and Shacklette 1975, Lagerwerff 1972, Shacklette and Boerngen 1984, Smart 1968, Vostal 1972, Wedepohl 1978). This element is very toxic to fungi and most plants as well as higher animals including man (Bowen 1966, Cook 1977, D'Itri 1972). Mercury ore deposites are found in geologically active belts, including the Pacific rim and a belt through Asia and the Mediterranean. The largest and richest deposits have been found in Spain (D'Itri 1972). Mercury is also known to be associated with many hydrothermal ore deposits of precious and base metals and has been used for geochemical prospecting for such deposits (Fleischer 1970, Oftedal 1940, McCarthy et al. 1969, Warren et al. 1966).

Annual global mobilization of mercury into the atmosphere has been estimated at 25,000 and 11,000 to 20,000 metric tons for natural and anthropogenic sources respectively (Galloway et al. 1982). Major sources of anthropogenic mercury include mining and smelting, manufacturing, combustion of fossil fuels, chlor-alkali plants, sewage disposal and agriculture (Blackwood et al. 1979, Bull et al. 1977, Cappon 1984, Crockett and Kinnison 1979, D'Itri

Table 3. Background copper levels in plants.

Medium	Use	Level (ppm DW)	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Shudu Sabbias	Dafa
	036		кезропае	raciiway	Receptor	Duracion	Method	Study Setting	Reference
U.S. Soils	Red Clover	(10.0) A	Background	Plant Uptake	Above Ground				
			-		Biomass	NR HNO3.	/Dithizone	Field	Kubota (1983)
U.S. Soils	Alfalfa	(8.8)	Background	Plant Uptake	H	NR	AAS	Field	Kubota (1983)
U.S. Soils	Alsike Clover	(8.3)	Background	Plant Uptake		NR	AAS	Field	Kubota (1983)
U.S. Soils U.S. Soils	Sweet Clover Ladino Clover	(7.9) (7.9)	Background	Plant Uptake		NR	AAS	Field	Kubota (1983)
U.S. Soils	Lotus	(7.4)	Background Background	Plant Uptake		NR NR	AAS AAS	Field	Kubota (1983)
U.S. Soils	Smooth Brome	(5.9)	Background	Plant Uptake Plant Uptake		NR NR	AAS	Field Field	Kubota (1983)
U.S. Soils	Bluegrass	(5.5)	Background	Plant Uptake	H	NR	AAS	Field	Kubota (1983)
U.S. Soils	Orchard Grass	(5.2)	Background	Plant Uptake		NR	AAS	Field	Kubota (1983) Kubota (1983)
U.S. Soils	Timothy	(4.6)	Background	Plant Uptake	H	NR	AAS	Field	Kubota (1983)
U.S. Soils	Fescue	(4.4)	Background	Plant Uptake	H	NR .	AAS	Field	Kubota (1983)
U.S. Soils	Wheatgrass	(4.0)	Background	Plant Uptake		NR	AAS	Field	Kubota (1983)
U.S. Soils	Broomsedge	(1.5)	Background	Plant Uptake	H	NR	AAS	Field	Kubota (1983)
Lateritic Gravelly	Trifolium			Parameter Administration			0.000	ALC 2 100 100	
Sand, pH 5.0	hirtum	5.3-12.3	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly Sand, pH 5.0	Ornithopus sativus	7.2-8.9	Background	Dlank Hataka		2 5	110		
Lateritic Gravelly	Pisum	7.2-0.9	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Sand, pH 5.0	arvense	7.4-8.4	Background	Plant Uptake	W	3-5 months	AAS	Field	61-4-t
Lateritic Gravelly	Lupinus		Duckground	rane opeake		3 3 monens	AAU	rieiu	Gladstones et al. (1975)
Sand, pH 5.0	consentinii	5.8-8.8	Background	Plant Uptake	H	3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	Ornithopus								Gradatones et al. (1975)
Sand, pH 5.0	compressus	7.0-8.1	Background	Plant Uptake	н .	3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly				•					010000000000000000000000000000000000000
Sand, pH 5.0	Sub. clover	6.2-10.7	Background	Plant Uptake	9	3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	110 00			700				91.0	South Control of Address of the Street Control of the Street Contr
Sand, pH 5.0	Alfalfa	5.1-7.6	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	Lupinus	4606	Dankanan d	Black Wataba	w	2.5	***	-1-1-	The second secon
Sand, pH 5.0 Lateritic Gravelly	luteus Vicia	4.6-8.5	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Sand, pH 5.0	atropurpurea	5.6-8.0	Background	Plant Uptake		3-5 months	AAS	Field	Cladeteres et al (1000)
Lateritic Gravelly	Lupinus	3.0-0.9	Background	Flanc Optake		3-3 months	AAO	rield	Gladstones et al. (1975)
Sand, pH 5.0	albus	3.2-6.8	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	Lupinus	(2 1/2) 2 1 E)		The second		- A WANTENIA			Gradatones et al. (1975)
Sand, pH 5.0	angustifolia	3.0-6.0	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	Arctotheca		2	10 × 10 × 10 × 10 × 10 × 10 × 10 × 10 ×					222223
Sand, pH 5.0	calendula	6.2-16.5	Background	Plant Uptake	•	3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly								AND THE PERSON NAMED	
Sand, pH 5.0	Rye	4.5-8.5	Background	Plant Uptake	•	3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	1 and a second s								County and the
Sand, pH 5.0	Wheat	3.3-5.6	Background	Plant Uptake	***	3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly Sand, pH 5.0	Barley	2.5-5.4	Background	Dlant Untaka		3-5 months	AAS	Field	AT 4.7
Lateritic Gravelly	Oats cv	2.5-5.4	Background	Plant Uptake		3-5 months	AAS	rield	Gladstones et al. (1975)
Sand, pH 5.0	Ballidu	2.4-6.8	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	Oats cv		David	Tranc opeane		3 3 110116110	mo		Gradacones et al. (1975)
Sand, pH 5.0	Avon	2.4-7.6	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	Bromus								01445 ct al. (1975)
Sand, pH 5.0	rigidus	3.9-9.7	Background	Plant Uptake	•	3-5 months	AAS	Field	Gladstones et al. (1975)
Lateritic Gravelly	Bromus			2 ESSORT 12 ESSOR 12 ESS	221	600 200 10078			
Sand, pH 5.0	mollis	3.8-6.6	Background	Plant Uptake		3-5 months	AAS	Field	Gladstones et al. (1975)
Piedmont Soils	Alfalfa	6.5-19.7	Background	Plant Uptake	,	Maturity	AOAC	Field	Price et al. (1955)
Piedmont Soils Piedmont Soils	Lespedeza Red Clover	6.0-14.2	Background	Plant Uptake		(July-Aug)	AOAC	Field Field	Price et al. (1955)
Piedmont Soils	Ladino Clover	10.5-29.0	Background	Plant Uptake	n		AOAC	Field	Price et al. (1955)
Piedmont Soils	Timothy	1.5-9.7	Background Background	Plant Uptake Plant Uptake	11	H	AOAC	Field	Price et al. (1955)
Piedmont Soils	Orchard Grass	8.0-18.5	Background	Plant Uptake			AOAC	Field	Price et al. (1955)
Cottenham Sandy Loam		3.9 W/MYC B	Background	Plant Uptake	Leaves	5 weeks	AAS	Field	Price et al. (1955) Gildon and Tinker (1983)
Cottenham Sandy Loan		2.8 WO/MYCB	Background	Plant Uptake	Leaves	5 weeks	AAS	Field	Gildon and Tinker (1983)
				Account of same			A	7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7	111NET (1783)

A () denotes means

B Mycorrhiza

Table 3. Background copper levels in plants, continued.

Medium		evel			xposur		N20 Y	20 2 4	and the M	Study	**********
ried I din	Use (pr	om DW)	Б	Response	Pathwa	Y	Receptor	Duration	Method	Setting	Reference
U.S. Soils	Cabbage (Wi)	20-150 (2	9) AWT	Background	Dlant	Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Corn (Ga)	70-150 (1	GG) AWT	Background		Uptake	Plant Uptake	NR NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Corn (Mo)	50-100 (7	G) AWT	Background		Uptake	Plant Uptake	NR NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Cucumber (Wi)			Background		Uptake	Plant Uptake	NR NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Potato (Wi)			Background							
U.S. Soils		100 200 ()	70 AWI	Background		Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Tomato (Ga)					Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Black Cherry			Background		Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
U.S. Soils	(Ga) Buckbush			Background	Plant	Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
	(Mo) 1	00-1500 (1	90) AWT	Background	Plant	Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Cedar (Mo)	20-200 (5	Ø) AWT	Background	Plant	Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Shagbark Hicko (Ky)	ry		Background		Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Black Oak (Ky)			Background		Uptake	Plant Uptake	NR NR		Field	Conner et al. (1976)
U.S. Soils	White Oak (Ky)		20) AWT	Background					AAS		
U.S. Soils	Smooth Sumac					Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
U.S. Soils	Winged Sumac			Background		Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
Sand Culture		50-200 (I	IW) AWT	Background		Uptake	Plant Uptake	NR	AAS	Field	Conner et al. (1976)
Sand Culture	Barley	12.4		Background		Uptake	Roots	9 weeks	CCM	Greenhouse/Nut. Soil	Agarwala et al. (1977)
	Barley	4.6		Background	Plant	Uptake	Young Leaves	9 weeks	CCM	Greenhouse/Nut. Soil	Agarwala et al. (1977)
Sand Culture	Barley	3.8		Background	Plant	Uptake	Old Leaves	9 weeks	CCM	Greenhouse/Nut. Soil	Agarwala et al. (1977)
and Culture	Barley	2.8		Background	Plant	Uptake	Stem	9 weeks	CCM	Greenhouse/Nut. Soil	Agarwala et al. (1977)
Solution Culture	Bush Beans	7.4		Background	Plant	Uptake	Leaves	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977d)
olution Culture.	Bush Beans	3.7		Background	Plant	Uptake	Stems	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977d)
ritish Columbia	Timber Milk			131 2 .16 10102502			Above Ground		20		
Soils	Vetch	2.8-9.0 (6.1)	Background	Plant	Uptake	Biomass	NR	AAS	Field	Fletcher and Brink (1969)
•	Arnica	4.5-8.5 (Background		Uptake	H	NR	AAS	Field	Fletcher and Brink (1969)
•	Pinegrass	4.5-10.2		Background		Uptake		NR	AAS	Field	Fletcher and Brink (1969)
" Ker	tucky Bluegrass	7-9-14-1	(9 9)	Background		Uptake		NR NR		Field	Fletcher and Brink (1969)
	Wheatgrass	3.7-7.9 (Background		Uptake			AAS	Field	Fletcher and Brink (1969)
SIT .	Lupine	7.9-8.5 (Background				NR	AAS	Field	Fletcher and Brink (1969)
olluted Soils	Agrostis gigantea	48-89	0.2)	Background near		Uptake		NR	AAS		
oil	Tall Fescue			Sudbury	Plant	Uptake	Shoots	Maturity	AAS	Field	Hogan et al. (1977)
,011	Tall rescue	3.3-10.2		Background (Penn)	Plant	Uptake	Above Ground			NATIONAL SERVICE	11 FeV 860 (00104 5044 0010)
rannia Much Cail-		The last of	u				Biomass	NR	NR	Field	Sopper and Seaker (1984)
rganic Muck Soils	Lettuce	(3.6) - (Background (Ont)		Uptake	Leaves	NR	AAS	Field	Czuba and Hutchinson (1980)
_	Lettuce	(18.4) -			Plant	Uptake	Roots	NR	AAS	Field	Czuba and Hutchinson (1980)
<u>.</u>	Celery	(4.6) - (Plant	Uptake	Leaves	NR	AAS	Field	Czuba and Hutchinson (1980)
	Celery	(13.8) -			Plant	Uptake	Roots	NR	AAS	Field	Czuba and Hutchinson (1980)
	Carrots	(4.8) - (6.5)	•		Uptake	Leaves	NR	AAS	Field	Czuba and Hutchinson (1980)
	Carrots	(6.5) - (17.0)			Uptake	Roots	NR	AAS	Field	Czuba and Hutchinson (1980)
**	Lettuce	(8.7)		2. H		Uptake	Leaf	Maturity	/	Field	Czuba and Hutchinson (1980)
**	Lettuce	(24.0)		**	Dlane	Uptake	Root	Autumn	AAS	Field	Czuba and Hutchinson (1980)
*	Celery	(7.8)								Field -	Czuba and Hutchinson (1980)
**	Celery	(12.6)				Uptake	Leaf/Stalk	*	AAS	Field	
•	Potato	(11.1)				Uptake	Root		AAS		Czuba and Hutchinson (1980)
9	Potato					Uptake	Leaf		AAS	Field	Czuba and Hutchinson (1980)
•	Carrot	(10.9)				Uptake	Roots/Tubers	W.	AAS	Field	Czuba and Hutchinson (1980)
**		(6.9)		- 2		Uptake	Leaf		AAS	Pield	Czuba and Hutchinson (1980)
•	Carrot	(4.9)				Uptake	Root	•	AAS	Field	Czuba and Hutchinson (1989)
	Parsnip	(8.9)		H		Uptake	Leaf	Spring	AAS	Field	Czuba and Hutchinson (1980)
•	Parsnip	(8.9)		H		Uptake	Root	Spring	AAS	Field	Czuba and Hutchinson (1980)
	Onion	(4.3)			Plant	Uptake	Leaf	Maturity, Autumn	/ AAS	Field	Czuba and Hutchinson (1980)
	Onion	(24.1)			Plant	Uptake	Root	10	AAS	Field	Czuba and Hutchinson (1980)
•	Onion	(3.7)		••		Uptake	Bulb		AAS	Field	Czuba and Hutchinson (1980)
	Cauliflower	(3.9)				Uptake	Leaf		AAS	Field	Czuba and Hutchinson (1980)
•	Cauliflower	(8.6)				Uptake	Root		AAS	Field	Czuba and Hutchinson (1980)
•	Cauliflower	(4.5)		•		Uptake	Flower Head		AAS	Field	Czuba and Hutchinson (1980)
•	Cabbage	(2.9)					Leaf			Field	Czuba and Hutchinson (1980)
	Cabbage	(7.7)				Uptake			AAS		Czuba and Hutchinson (1980) Czuba and Hutchinson (1980)
				0.000	PIANT		Root		AAS	Field	

Table 3. Background copper levels in plants, continued.

Medium	Use	Level (ppm DW)	Hazard Response	Pathway	Receptor	Duration	Method	Study Setting	Reférence
Plainfield Loamy	98	pc.	Background	• 6		Pirst Trifo	-		
Sand	Snap Beans	16.7-18.3	•	Plant Uptake	Leaves	liate Leaf		Field	Walsh et al. (1972)
	Snap Beans	8.3-24.7		Plant Uptake	Leaves	Pod Set	AAS	Field	Walsh et al. (1972)
•	Snap Beans	14.3-17.0		Plant Uptake	Leaves	Maturity	AAS	Field	Walsh et al. (1972)
	Snap Beans	10.7-16.0		Plant Uptake	Stems	Maturity	AAS	Field	Walsh et al. (1972)
•	Snap Beans	12.0-18.0		Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
olution Culture	Per. Ryegrass	3.8	Background	Plant Uptake	Shoots	21 Days	AAS	Greenhouse	Jarvis (1978)
rilsham Loam	Per. Ryegrass	8.8	Background	Plant Uptake	Shoots	42-102 days	AAS	Greenhouse	Jarvis (1978)
rilsham Loam	Per. Ryegrass	12.6	Background	Plant Uptake	Roots	102 days	AAS	Greenhouse	Jarvis (1978)
itzville Silt Loam	Soybeans	3.90	Background	Plant Uptake	Tops	60 days	AAS	Greenhouse	Cataldo and Wildung (1978)
ubbard Coarse Sand	Snap Beans	4.1	Background (Unfert)	Plant Uptake	Pods	Maturity	AAS	Field	Latterell et al. (1978)
•	Snap Beans	2.8	Background (Fert)	Plant Uptake	Pods	Maturity	AAS	Field	Latterell et al. (1978)
•	Snap Beans	10.0	Background (Unfert)	Plant Uptake	Leaves	Early Bloom			
	•		-			Stage	AAS	Field	Latterell et al. (1978)
•	Snap Beans	8.2	Background (Fert)	Plant Uptake	Leaves	•	AAS	Field	Latterell et al. (1978)
.S. Soils	Lettuce	1.6-18.3 (6.3)	Background	Plant Uptake	Edible Portions	NR	ICP	Field	Wolnik et al. (1983)
.S. Soils	Peanuts	0.91-22 (8.6)	Background	Plant Uptake	**	NR	ICP	Field	Wolnik et al. (1983)
.S. Soils	Potatoes	0.73-14 (5.0)	Background	Plant Uptake		NR	ICP	Field	Wolnik et al. (1983)
.S. Soils	Soybeans	3.5-29 (12)	Background	Plant Uptake	11	NR	ICP	Field	Wolnik et al. (1983)
.S. Soils	Wheat	2.5-9.9 (5.0)	Background	Plant Uptake		NR	ICP	Field	Wolnik et al. (1983)
.S. Soils	Sweet Corn	0.89-4.3 (2.1)	Background	Plant Uptake		NR	ICP	Field	Wolnik et al. (1983)

Table 4. Elevated copper levels in plants.

Medium	Use	Level (ppm DW)	Hazard Response	Exposure Pathway	Receptor	Duration	Method ·	Study Setting	Reference
Sand Culture									
Frilsham Loam	Barley Perennial	2120	31 % YR	CuSO ₄ 5H ₂ O	Roots	6 weeks	cc	Greenhouse	Agarwala et al. (1977)
III Sham Loam	Ryegrass	377.3	49 % YR	Plant Uptake	Name -	102 days	AAS	Greenhouse	Jarvis (1978)
and Culture	Barley	156	34 % YR	Cu (NO ₃) ₂ 3H ₂ O Cu SO ₄ 5H ₂ O	Old Leaves	6 weeks	CC	Greenhouse	Agarwala et al. (1977)
Varsaw Sandy Loam	Corn	127.8	60 % YR	Amended Sludge		0 WEEKS	CC	Greennouse	Agarwara ec ar. (1577)
		127.0	00 6 TK	Amended brudge	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a
Frilsham Loam	Perennial								
	Ryegrass	94.7	7.8 % YR	Plant Uptake	Roots	102 days	AAS	Greenhouse	Jarvis (1978)
Sand Culture	Oats	92	Specific Cu Tox. 44 %				Spectro-		
Warsaw Sandy Loam		02.0	Reduction in Plant Height		Shoots	40 days	chemical	Greenhouse	Hunter and Vergnano (195
varsaw sandy Loam	Corn	83.8	41 % YR	Amended Sludge		6 marks	AAS	-	Cunningham et al. (1975a
Warsaw Sandy Loam	Corn	56.1	45 % YR	Amended Sludge	Biomass	6 weeks	AAS	Greenhouse	Cunningnam et al. (1975a)
arsaw bandy boam	COLII	30.1	45 6 IK	Amended Studge	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a
Warsaw Sandy Loam	Rye	53.8	39 % YR	Amended Sludge		o weeks	nno	Greennouse	Cumingham et al. (1575a)
Juney Louis	w.y.c	33.0	39 6 IR	Amended Studge	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a
Sand Culture	Barley	52	34 % YR	CuSO4 5H20	Stem	6 weeks	CC	Greenhouse	Agarwala et al. (1976)
IR	Cabbage	50	Reduced Yield	NR	NR	NR	NR	NR	Hara and Sonoda (1979)
and Culture	Barley	49	34 % YR	CuSO4 5H20	Young Leaves	6 weeks	CC	Greenhouse	Agarwala et al. (1976)
larsaw Sandy Loam	Corn	43.7	59 % YR	Amended Sludge					The second of th
					Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a
Sandy Soil"	White Clover		50 % YR	Plant Uptake	Shoots	8 weeks	AAS	Greenhouse	Dijkshoorn et al (1979)
and Culture	Oats	37	Highly Chloratic 13 %				Spectro-		
		88.8	Reduction in Plant Height		Shoots	40 days	chemical	Greenhouse	Hunter and Vergnano (195
arsaw Sandy Loam	Corn	35.1	51 % YR	Amended Sludge				4	
					Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a
olo Loam	Bush Bean	34.3	83 % YR	Plant Uptake	Leaves	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
larsaw Sandy Loam	Rye	30.9	29 % YR	Amended Sludge		(0.0200040.020)		-content-conservation	0
Varsaw Sandy Loam	Corn	30.5	44 9 00	>	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a
aroun banay boam	COLII	30.3	44 % YR	Amended Sludge	Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a
olo Loam	Bush Bean	28.8	26 % YR	Plant Uptake	Leaves	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
lainfield Loamy San		27.3	84 % YR	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
arsaw Sandy Loam	Rye	26.1	14 % Yield Increase	Amended Sludge			11012		Manager and Manager
· · · · · · · · · · · · · · · · · · ·	20 4 (\$20)	12.7.5			Biomass	6 weeks	AAS	Greenhouse	Cunningham et al. (1975a
rilsham Loam	Perennial								PETTONONITRON → PRINCE MEDICAL DELEVATOR AND PRINCESSOR
	Ryegrass	25.5	6.9 % YR	Plant Uptake	Shoots	102 days	AAS	Greenhouse	Jarvis (1978)
Solution Culture	Perennial								
	Ryegrass	24.7	3.4 % YR	CuSO ₄ 5H ₂ O	Shoots	21 days	AAS	Greenhouse	Jarvis (1978)
lainfield Loamy San		20.7	38 % YR	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
olo Loam	Bush Bean	20.3	69 % YR	Plant Uptake	Stem	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
lainfield Loamy San	d Smap Beans	20.0	17 % YR (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
lainfield Loamy San	d Snap Beans	19.6	34 % YR	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
lainfield Loamy San and Culture		18.7	5 % YR (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
Sandy Soil"	Barley Perennial	18-21 (20)	10 % YR	Plant Uptake	Shoot	5 leaf stage	AAS	Greenhouse	Davis and Beckett (1978)
Saucy Soll	Ryegrass	18	50 % YR	Dlant Untako	Shoot	4 weeks	AAS	Greenhouse	Dijkshoorn et al (1979)
olo Loam	Bush Bean	17.8	12 % YR (N.S.)	Plant Uptake Plant Uptake	Leaves	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
lainfield Loamy San		17.7	24 % YR (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
lainfield Loamy San	d Snap Beans		6.5% Yield Increase (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
lainfield Loamy San	d Snap Beans	17	4 % YR (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
and Culture	Oats	17	Normal	CuSO ₄ 5H ₂ O	Shoots	40 days	Spectro-		
		100	Constant Control		101111717777		chemical	Greenhouse	Hunter and Vergnano (19
lainfield Loamy San	d Snap Beans	16.3	14 % YR (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
rilsham Loam	Perennial		- c cos wasar as	Plant Uptake	and and a second a se				
	Ryegrass	15.3	7.8 % YR	Cu (NO3) 2 3H20	Shoots	42-102 days	AAS	Greenhouse	Jarvis (1978)
lainfield Loamy San		15	76 % YR	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
lainfield Loamy San		15	1.6 % YR (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
Sandy Soil"	Plantain	15	50 % YR	Plant Uptake	Shoots	6 weeks	AAS	Greenhouse	Dijkshoorn et al (1979)

Table 4. Elevated copper levels in plants, continued.

Medium	Use	Level	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Prilsham Loam	Perennial			Plant Uptake					
	Ryegrass	14.7	49 % YR	Cu (NO3) 2 3H20	Shoots	42-192 days	AAS	Greenhouse	Jarvis (1978)
olution Culture	Perennial								
	Ryegrass	14.2	17.6% Yield Increase	CuSO ₄ 5H ₂ O	Shoots	21 days	AAS	Greenhouse	Jarvis (1978)
lainfield Loamy San	d Snap Beans	13.7	21 % YR (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
lainfield Loamy San	d Snap Beans	12.3 1.	.5 % Yield Increase (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
lainfield Loamy San	d Snap Beans	12.3	3 % YR (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
olo Loam	Bush Bean	11.7	14 % YR (N.S.)	Plant Uptake	Stem	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
lainfield Loamy San	d Snap Beans	11	26 % Yield Increase	Plant Uptake	Pods	Maturity	AAS	Pield	Walsh et al. (1972)
colution Culture	Perennial					9 75			
	Ryegrass	10.7	13 % Yield Increase	CuSO4 5H20	Shoots	21 days	AAS	Greenhouse	Jarvis (1978)
lainfield Loamy San	d Snap Beans	10.3	15 % Yield Increase (N.S.)	Plant Uptake	Pods	Maturity	AAS	Field	Walsh et al. (1972)
olo Loam	Bush Bean	10	17 % YR (N.S.)	Plant Uptake	Leaves	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
olo Loam	Bush Bean	9.5	0.8 % YR (N.S.)	Plant Uptake	Stem	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
olution Culture	Perennial								And the state of t
	Ryegrass	5.9	7.3 % Yield Increase	CuSO ₄ 5H ₂ O	Shoots	21 days	AAS	Greenhouse	Jarvis (1978)
olo Loam	Bush Bean	5.0	1.1 % YR (W.S.)	Plant Uptake	Stem	17 days	ES	Greenhouse/Soil Pots	Wallace et al. (1977a)
and Culture	Soybeans	3.22	27 % YR	Cu Solution	Shoots	42 days	AAS	Greenhouse	Vesper and Weidensaul (1978
Sand Culture	Soybeans	3.15	11 % YR	Cu Solution	Shoots	42 days	AAS	Greenhouse	Vesper and Weidensaul (1978

1972, Lindberg and Turner 1977, Lindberg et al. 1979). Typical condenser stack gas emissions from primary lead smelters of Ø.12 Kg mercury per metric ton of lead ore, have been reported (Blackwood et al. 1979). Agricultural use of mercury for seed treatments and fungicides has decreased in recent years due to environmental concerns and the danger of mercury entering the food chain (Friberg and Vostal 1972).

Mercury is found in three valence states: metallic Hg, Hg⁺ and Hg⁺⁺ and forms hundreds of inorganic and organic compounds (Battelle 1977). Mercury has received very little attention concerning the existence and levels of specific chemical forms which could influence the soil chemistry and eventual plant uptake (Cappon 1984). The three common forms of mercury; elemental, inorganic salts, and organic compounds, are all toxic but the organic compounds, especially the alkyl mercury compounds, appear the most hazardous (Blackwood et al. 1979). The toxicity of mercury to terrestrial plants apparently depends more on chemical form than on its concentration (Ratsch 1974).

2.2.1 Total mercury levels in soils

Mercury is immobilized in soils through three basic processes: 1) formation of relatively insoluble forms such as HgS, metallic Hg and Hg_2^{2+} ; 2) sorption by soil colloids, especially clays; and 3) complexation by organic ligands (Gilmour and Miller 1973, Hogg et al. 1978, Kabata-Pendias and Pendias 1984, Lindberg et al. 1979, Weaver et al. 1984). Accumulation of mercury in soils is controlled by organic complex formation and by precipitation (Kabata-Pendias and Pendias 1984). Mercury is usually retained in soils as slightly mobile organocomplexes (Kabata-Pendias and Pendias 1984) and organic soils generally have elevated mercury levels compared to mineral soils (Chattopadhyay and Jervis 1974, Frank et al. 1976). Lindberg et al. (1979) reported a control soil (pH 5.6) in the vicinity of the Almaden mercury deposit in Spain in which organo-clay complexes contained 38 percent of the soil mercury distribution, and the clay-mineral fraction and mineral fraction contained 35

percent and 6 percent of the soil mercury respectively. Dudas and Pawluk (1977) found a significant relationship between mercury levels and 1) cation exchange capacity (CEC), 2) organic matter content and 3) exchangeable calcium only in poorly drained soils. These authors found no significant relationships between soil mercury levels and pH, organic matter content, CEC or exchangeable calcium for well drained or solonetz soils. This may have been partially due to the low mercury levels found in this study (only one sample >0.060 ppm mercury).

Mercury levels in soils may be decreased by three general mechanisms: 1) volatilization, 2) leaching, and 3) plant uptake. Mercury in soils is unique in that it is one of the few metals that is readily volatilized and lost to the atmosphere from surface soils (Frear and Dills 1967, Gilmour and Miller 1973, Lindberg et al. 1979). Estimates of mercury volatilization range from 10 to 32 percent and 44 to 56 percent by Hogg et al. (1978) and Gilmour and Miller (1973) respectively. Lindberg et al. (1979) measured $\emptyset.13$ and $\emptyset.33$ ug mercury/m² volatilization per hour at 25°C from background soils and soils near the Almadin mercury mine respectively. Hogg et al. (1978) determined 10 to 32 percent of all applied soil mercury in their experiments was lost presumably by volatilization. The formation of volatile mercury has been shown to be stimulated by increased soil moisture, pH, and temperature (Frear and Dills 1967, Gilmour and Miller 1973).

Leaching of mercury from surface soils is limited but apparently does occur. Significant movement of spiked mercuric chloride has been demonstrated from the Ø to 10 cm soil layer to the 20 to 30 cm soil layer and 30 to 40 cm soil layer for loam and loamy sand soils respectively (Hogg et al. 1978). However, these authors found the movement of mercuric chloride, phenylmercuric acetate (PMA), and methyl mercuric chloride (MMC) to be "severely limited" even in light textured soils of low organic content. No statistical significance between mercury levels in A and C horizons of 16 Manitoba soil series (loamy through clay soils) was found by Mills and Zwarich (1975). Mercury levels in

rangeland soils of the Powder River Basin are only slightly higher in subsoils than in surface soils (Connor et al. 1976). Most mercury deposited on soil is confined to the upper 3 cm (Battelle 1977) and is usually present in surface soils at several times the levels in subsoils (Kabata-Pendias and Pendias 1984).

Mercury removal from the soil system by plant uptake is also limited. Very high mercury concentrations have been found in roots of several species of plants. Roots of rice plants were reported to contain 1000 ppm mercury while the grain from these plants contained only 0.5 ppm mercury (Ishizuka and Tanaka 1962). Roots of most grain crops and many vegetable crops remain in the soil and hence, much of the mercury uptake by plants remains in the soil system. Only the amount translocated to the above ground biomass is readily removed (Section 2.3.2).

The predominate form of inorganic mercury is likely to be $Hg (OH)_2$ at $pH \geq 7$ (Kabata-Pendias and Pendias 1984). Acid gley soils may contain HgS (cinnibar) or metallic mercury (Kabata-Pendias and Pendias 1984). Methylation of inorganic mercury has been confirmed in agricultural soils (Cappon 1984). Methylmercury formation in soils is apparently directly proportional to clay and soil organic content, moisture content, temperature and substrate mercury concentrations (Cappon 1984). The process is enhanced by pH levels <6.0.

The effect of soil mercury on mycorrhizal activity is uncertain but it has been shown that additions of zinc, copper, nickel or cadmium can adversely affect mycorrhizal fungus and thus the phosphorus nutrition of the host (Gildon and Tinker 1983). Given the known toxicity of mercury and its use as a fungicide, it is quite likely that a similar response would occur from elevated soil mercury levels. This subject needs further research.

Mercury levels in the earth crust or lithosphere have been estimated from 0.05 to 0.5 ppm (Jenkins 1980, Swaine 1955). Typical values reported for sandstones, shales and carbonates are 0.030 ppm, 0.400 ppm and 0.040 ppm respectively (Wedepohl 1978)

(Table 5). Typical soil mercury levels are reported to range from 0.03 to 0.8 (Bowen 1966, Swaine 1955). Ratsch (1974) reported a United States soil mercury range of 0.10 to 0.500 ppm and most background soil mercury levels found in the literature fall within this range (Table 5). Frank et al. (1976), Mills and Zwarich (1975) and Dudas and Pawluk (1977) determined background mercury levels in soils of Ontario, Manitoba and Alberta respectively. The maximum range for these provinces was $\emptyset.01$ to $\emptyset.78$ ppm mercury with a maximum mean value of 0.41 ppm mercury found in an Ontario organic soil. Mean values for all Ontario and Manitoba soils tested were 0.07 and 0.033 ppm respectively. only background mercury soil level exceeding 0.78 ppm was associated with organic muck (gley) soils in which the mercury content ranged from 1.97 ppm at the soil surface to 1.20 ppm at the 22.5 to 30 cm depth increment (Chattopadhyay and Jervis 1974). Background mercury levels for surface soils in the Powder River Basin of Montana and Wyoming were reported to range from $\emptyset.01$ to 0.04 ppm with a geometric mean of 0.020 (Connor et al. Background surface soil sample sites near the Helena Valley had a reported mercury range of 0.06 to 0.12 ppm with a mean value of 0.08 (EPA 1986).

Elevated soil mercury levels have been documented near industrial sites and urban areas (Table 6). Carey et al. (1980) found significant differences between urban and suburban soils for 5 midwestern and eastern cities in the United States. The absolute values derived in this study were questionable due to the soil drying methods employed. Klein (1972) found mercury levels in soils of the Grand Rapids Michigan area of 0.10 and 0.11 ppm for residential and agricultural soils respectively. This author also reported to 0.14 and 0.33 ppm mean soil mercury concentrations for industrial sites and an airport respectively. Klein and Russell (1973) found increased soil mercury levels (0.0079 versus 0.00102 ppm) near a coal fired power plant in Michigan.

Elevated soil mercury levels have been found near many smelters and ore deposites (Heilman and Ekuan 1977, Lindberg et

Table 5. Background total mercury levels in soils.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	r Duration	Method	Study Setting	Reference
Biotite	Not Given	0.01-0.38	Background	NR	NR	Not Applicable	NR	Field samples	Wedepohl et al. (1978)
Garnet	m	0.006-0.007	•	NR	NR	n	NR	"	, н
Staurolite	•	0.005-0.963		NR	NR	•	NR	н	
Sillimanite	n	0.002-2.535	•	NR	NR	**	NR	"	
Shales	11	0.400	•	NR	NR	**	NR		
andstones	**	0.030	•	NR	NR	**	NR		•
arbonates	**	0.040	•	NR	NR	11	NR	,	•
oal (Bit)	Fuel	0.008-0.022	**	NR	NR	*	NR		
oal (Brown)	Fuel	0.001-0.025	•	NR	NR	**	NR	Ti .	
ithosphere	Not Given	0.5		NR	NR	311	NR	NR	Swaine (1969)
oils	u	0.03	••	NR	NR	n	NR	NR	•
"		0.01-0.3		NR	NR	w	NR	NR	Bowen (1966)
"	н	(0.071)	.W .	NR	NR	**	NR	Field	Shacklette an Boerngen (198
oils, U.S.	"	0.010-0.500	н	NR	NR	•	FLAAS		Ratsch (1974)
arth Crust		0.05	H	NR	NR	***		m.	Jenkins (1986
oil, pH 7.4	Garden	0.156		Plant uptake	Humans		GLC	W.	Cappon (1984)
ncultivated Soil	Not Given	0.045-0.160	•	!!	NR			н	Erdman et al. (1976)
nmineralized CA soils		0.02-0.06	*	NR	NR	: H •		,	
Soil	**	0.01	**	NR NR	NR NR	,,	NR	<u>"</u>]	Pleischer (1976
" (Japan)		0.28	•	Plant uptake	NR		NR NR	и .	EPA (1984) Kitagishi and Yamane (1981

Table 5. Background total mercury levels in soils, continued.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor Du	ration	Method	Study Setting	Reference
Muck soil (Sur)	Market Garden	1.97	Background	Plant uptake	Humans Not	Applicable	IPAA	Field	Chattopadhyay and Jervis (1974)
		1.18	11	: **	,		•	**	•
" 0-7.5 cm " 7.5-15 cm	**	1.62		**	•		11	• •	
" 15-22.5 cm	**	1.80	11	•	**	99	"	•	
" 22.5-30 cm	**	1.20	**	"		11		**	
Soil, 0-4 inch	Crops/Range	(0.08)	H		Crops/ forage	•	EPA CV		EPA (1986)
elena Valley p	18.0 "	0.06-0.12	•	**		u		•	
ur Soil/Powder	Native Range	(0.020) Geometric	c "	. "	Sagebrush	<u> </u>	AAS		Connor et al. (1976)
iver Basin		0.01-0.04			**	11	·		
11/2003		(0.023) Geometri	c "	,,	n	***	***		**
ubsoil/Powder iver Basin		0.01-0.04		"		u	•	11	"
oils UK	Agriculture	0.04-0.19 (0.106) "	m	Grass	u i	H	311	Bull et al. (1977
lew York Soils	Orchards	Ø.2	n		Garden M. Plants	aturity	FLAAS	n	Elfving et al (1978)
н		0.6	None Noted	n	•	n	111	H	n
.6 Manitoba soi	ls Crops	0.020-0.053	Background	**	Grains/ No forage	t applicabl	.e "	н	Mills and Zwarich (1975
н	•	(0.033)	**	M	11			;11	W)
ll Winnipeg Urba	an "	(0.029)			NR	n ,	••	"	n
.26 Ontario soi	ls Field Crops	0.01-0.70 (0.07)			Small grains/ forage	•	FLAAS - CV	"	Frank et al. (1976)
Ontario Mineral Soils	Vegetables	Ø.02-0.78 (0.10)		w	Vegetables	**	u	u	н
Ontario Organic Soils	n	Ø.Ø5-1.11 (Ø.41)		**		***	n.	,	n

Table 5. Background total mercury levels in soils, continued.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor D	uration	Method	Study Setting	Reference
Ontario Soils	Orchards	0.03-1.14 (0.29)	Background	Plant uptake	Apples Not	Applicable	FLAAS - CA	Field	Frank et al. (1976)
10.	H	0.02-0.27 (0.07)	W	II .	Cherries	н	n	m '	
•	#	0.02-0.18 (0.06)	11	··	Peaches	111		m	•
	•	0.04-0.31 (0.10)		u ·	Grapes	m	***		*
Ontario Sandy Soils	Crops	0.01-0.70 (0.06)	W .	u	NR		**		u.
Ontario Loam Soils	**	0.02-0.78 (0.09)		и -	NR	n	*	w	n
Ontario Clay Soils	•	0.03-0.46 (0.08)	w .	W	NR	n	•	NI NI	w
Alberta Brown Soil pH 7.2		0.024 <u>+</u> 0.007	" (Well drain	ned)	Cultivated Crops	10	FLAAS		Dudas and Pawluk (1977
Alberta Black Soil pH 6.4	*	0.027 <u>+</u> 0.008	**		n	п	u		
Alberta Gray Soil pH 6.5		0.024 <u>+</u> 0.005		и	n	30	11	**	
Alberta Brown Soil pH 6.5	u u	0.023 <u>+</u> 0.002	" (Poorly dra	" ained)	n	, m	11	11	"
Alberta Black Soil pH 6.9		0.035 <u>+</u> 0.015	•			n		11	91
Alberta Gray Soil pH 7.4	u	0.037 <u>+</u> 0.011	••	"		n n		n	tt
Alberta Brown Soil pH 6.4	n	0.016 <u>+</u> 0.003	" (Solonetz)	u	W	**		w 4	. G

Table 5. Background total mercury levels in soils, continued.

		Level (ppm DW)	Hazard	Exposure				Study	
	Use	means in ()	Response	Pathway	Receptor	Duration	Method	Setting	Reference
Medium	USE	means in ()	кезропае	14411144			457.7007.70		
Berrien Co., M	I			2000A00 10 10 10 10				-7-33	ab 1-1 - 4 b - /100a
pH (6.6)	Orchards	0.059-0.68 (0.23)	Background	Plant uptake		Maturity	FLAAS	Field	Shacklette (1980
pH (5.4)	**	0.031-0.078 (0.044)			Pears	"	***		***
Wayne Co. NY						11	* *		~
pH (5.5)	Orchards	0.14-0.32 (0.20)	90		Apples	"			
pH (5.5)	"	0.040-0.085 (0.059)		**	Peaches		"	11	
рн (6.6)	31	0.047-0.096 (0.060)			Pears	W			
pH (6.6)	**	0.04-2.6 (0.15)	11	**	Plum	"	•	••	
Yakima Co. WA							11		
pH (6.6)	Orchards	0.018-0.11 (0.044)	11		Apples	TH.	"	XX	*
pH (7.9)	***	0.01-0.16 (0.030)	ii .	"	European		***	10	
P					Grapes	III	"		
pH (5.7)	**	0.032-0.063 (0.043)	11	"	Peaches	**		11.	
pH (6.3)	n	0.19-0.040 (0.29)	11	**	Pears	"			<u></u>
pH (6.8)		0.010-0.037 (0.025)		11	Plum	"			
pH (7.1)	Field Crops	0.026-0.041 (0.032)	H .	**	Potatoes	12	***		
рн (6.6)	11	0.03-0.67 (0.046)	"	**	Tomatoes	**	29.		
Gloucester Co.	NJ							,,	
pH (5.5)	Orchards	0.01-0.13 (0.071)	**	n	Apples	**	***		
Mesa Co. CO								**	
pH (7.8)	31	0.023-0.065 (0.041)	**	11	Apples	**	11	,,	
pH (7.7)		0.026-0.058 (0.040)	11	11	Peaches		***	. "	
pH (8.0)	***	0.019-0.20 (0.042)		11	Pears	10	**	,,	
pH (7.6)		0.029-0.062 (0.040)	11	11	Plum		**		
pH (7.9)	Field Crops	0.029-0.046 (0.036)	· 11		Dry Beans	; "	"		•
Twin Falls Co.							100		
pH (8.1)	11	0.03-0.046 (0.038)	11	***	Dry Beans	3 "		••	-
pH (8.2)	11	0.023-0.037 (0.031)		**	Potatoes	**	11	,11	-
pH (8.3)	Vegetables	0.030-0.052 (0.037)		"	Snap Bear	ns "	"	300	
pH (8.0)	109000000	0.024-0.043 (0.035)		11	Sweet Cor	en "	"	11	3.10
San Joaquin Co	CA	a saga nanan sanan .			European				
pH (6.4)	Orchards	0.01-0.039 (0.021)) **	"	Grapes	99	- 10	11	
pH (6.8)	"	0.030-0.043 (0.035		99	Peaches	**	W .	**	
pH (7.0)	II .	0.057-0.10 (0.073)	"	n	Pears	**	**	11	
	Venetables	0.043-0.13 (0.082)	**	11	Cucumbers	s "	11	. 11	
pH (7.5)		0.016-0.035 (0.026	, "	°M.	Dry Beans	s "	11	11	"
рн (7.0) рн (8.5)		0.010-0.039 (0.026		ii = 1	Tomatoes	***	**		

2

Table 5. Background total mercury levels in soils, continued.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Alberta Black Soil pH 5.7	Crops	Ø.028 <u>+</u> Ø.007	Background Poorly Drain	Plant uptake	Cultivat crops	ed Not Applicable	FLAAS	Field	Dudas and Pawluk (1977)
Alberta Gray Soil pH 6.2	W · .	Ø. Ø41 <u>+</u> Ø. Ø29				**	"		
Canadian Soils	Not Given	0.005-0.11(0.05	9) "	NR	NR	**	NR	Field	McKeague et al (1979)
	Uncultivated	0.06	Ħ	NR	NR		AAS	m "	McKeague and Wolynetz (1980)

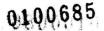
Table 6. Elevated total mercury levels in soils.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Arenosa Fine Sand pH 4.7	Pasture	50	Toxic	Plant uptake	Bermuda grass	6 weeks	HgCl ₂ Added	Field/soil pots	Weaver et al. (1984)
Weswood Silt Loam pH 7.7	Pasture	8	Reduced Plant Growth	. " "	n	6 weeks	n		
Houston Black clay pH 7.6	Pasture	50	Non Toxic	n 11	n	6 weeks			n
Polluted Soils	Ø.:	24-0.40(0.36)	Not Noted	99 99	Barley	N/A	INAA	Field	Singh and Steinnes (1976)
Hazelwood Silt Loam pH 5.1	Vegetables/Oat	s 20	Yields not stated	" "	Roots/ Leaves/ Grain/ Tubers/ Pods/Vi		HgCl ₂ Added	Greenhouse soil pots	John (1972)

al. 1979, McCarthy et al. 1970, Ratsch 1974, Shacklette 1970, Warren et al. 1966). Ratsch (1974) reported up to 11 ppm mercury in garden soils near the Ruston copper smelter in Tacoma, Washington. The hazard evaluation of excess total soil mercury is discussed in Section 3.2.

2.2.2 Mercury levels in vegetation

An increase in plant uptake of mercury with increased soil mercury levels has been demonstrated (Kabata-Pendias and Pendias 1984), however, plant translocation of soil mercury is low. Lead, chromium and mercury are so strongly held in root cells that very little is translocated to shoots of crop plants (Chaney 1984). Hogg et al. (1978) found fine roots of brome grass contained 43 to 102 ppm mercury, a level 2 to 4 times higher than levels found in primary or secondary roots. Similar results have been reported for rice, in which 1000 ppm was found in the roots and 0.5 ppm in the grain (Ishizuka and Tanaka 1962). Direct uptake of atmospheric mercury by alfalfa plant leaves has been suggested by Lindberg et al. 1979. Limited data suggest this mechanism is present in other species (Hitchcock and Zimmerman 1957). Although mercury translocation within plants is low, it is significant. John (1972) found elevated levels in edible portions of many vegetables grown on mercury ammended soil. author reported radish tubers and spinach leaves accumulate the highest levels at 0.695 and 0.663 ppm respectively. Translocation of mercury to grains is apparently limited. Dudas and Pawluk (1977) found mercury levels in wheat, barley and oat straw to be 2 to 5 times higher than in the respective grains. Translocation of methylmercury from seed dressings to the first generation of wheat and peas has also been demonstrated (Kabata-Pendias and Pendias 1984, Lagervall and Westoo 1969, Smart 1968). The mercury concentration in some plant materials is apparently enhanced by lower temperatures and a reduced photoperiod. et al. (1978) found the mercury concentration of bromegrass increased as the photoperiod and temperatures decreased during the autumn months.



The toxicity of mercury to plants is caused by the affinity of mercury to sulfhydryl groups and the resulting disruption of metabolic processes (Kabata-Pendias and Pendias 1984). The high concentrations of mercury observed in roots inhibits potassium uptake (Kabata-Pendias and Pendias 1984). The symptoms of mercury poisoning in plants are usually manifested in stunting of seedling growth, decreased root mass, and inhibition of photosynthesis (Kabata-Pendias and Pendias 1984).

Background mercury levels in plants have been relatively well defined (Table 7). The typical range is from trace amounts (0.001 ppm) to about 0.200 ppm mercury. The highest mercury background level that has been reported in the literature reviewed is 0.237 for radishes (John 1972). Nearly all edible vegetative products contain <0.100 ppm mercury.

Considerable variation is apparent among different plant species in their uptake of mercury under elevated conditions (Table 8). Small grain cereal crops exhibit small but significant increases in grain mercury contents. John (1972) found the mercury content of oat grain to increase from the 0.009 ppm background level to 0.020 ppm in plants grown in soil with 20 ppm mercury content. Radish tubers, grown under the same conditions, increased from 0.013 to 0.663 ppm mercury, an increase of over 50 times. The limited amount of phytotoxic mercury plant concentration data derived from reviewed literature suggests a wide phytotoxic range occurs (from 0.2 ppm to 6.4 ppm), dependent on many factors including the mercury compound, the experimental design and the plant species. These problems and hazard level selection are discussed in Section 3.2.

2.3 Selenium Levels in Soils and Plants

Selenium is an element commonly found in trace quantities throughout the ecosystem. Selenium is not regarded as an essential element for most crop plants, but some indicator species have been shown to respond to selenium uptake (NRC 1976). This element has an important role in animal nutrition and disease. Selenium in livestock diets is required in minute amounts to

Table 7. Background mercury levels in plants.

1edium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
egetation	Sweet Corn	0.003, 0.0046 H	Background	Plant uptake	Grain	NR	NR	NR	Kabata-Pendias and Pendias (1984)
egetation	Bean	0.003, 0.011	**	W s	Pod	NR	NR	NR	rendias (1964)
egetation	Bean	0.017 WW, 0.07 WW	**	10	100	NR	NR NR	NR	11
egetation	Carrot	0.086, 0.0057	**		Root	NR	NR NR	NR	11
'egetation	Lettuce	0.0083	***	**	Leaves	NR	NR	NR	H
egetation	Lettuce	<0.0006 WW	11	Ü	neaves	NR	NR NR	NR NR	
egetation	Cabbage	0.0065	**		**	NR NR	NR NR	NR NR	
egetation	Cabbage	0.010 WW	11	**	11				
egetation	Beet	0.003 WW	11	***	Root	NR NR	NR NR	NR NR	Kitagishi ar Yamane (1981 Kabata-Pendias and
egetation	Potatoes			₩				573. E.S.	Pendias (1984)
egetation		0.047, <0.010			Tuber	NR	NR	NR	"
egetation	Potatoes	0.003 WW, 0.12 W	₩ "		190	NR	NR	NR	**
	Onion	<0.010	55A		Bulb	NR	NR	NR	
egetation	Onion	0.007 WW	**	11	**	NR	NR	NR	
egetation	2	0.001 WW, 0.011 WV		11	Unpeeled Fruit	NR	NR	NR	
egetation	Tomato	0.0031, 0.034	**	••	Fruit	NR	NR	NR	**
egetation	Tomata	0.001 WW	w	n	11	NR	NR	NR	**
egetation	Apple	<0.010	m .	**	**	NR	NR	NR	
egetation	Apple	0.010 WW	**	11		NR	NR	NR	11
egetation	Orange	0.0026	**	**	111	NR	NR	NR	•
egetation	Lemon	0.043 WW	**		11	NR	NR	NR	11
egetation	Mushrooms	0.0035	н		Caps, Stalks	NR	NR	NR	**
egetation	Green/Yellow Vegetables	0.02 WW	n		Edible Parts	NR	NR	NR	Kitagishi a Yamane (198)
egetation	Lettuce	0.031	n	**	Leaves	35 days	HNO3/HC104/	Greenhou	
egetation	Lettuce	0.112	**	**	Roots	"	FLAAS	Soil pot	
egetation	Spinach	0.094	11	••	Leaves	55 days	**	, POC	
egetation	Spinach	0.095		91	Roots	55 days	**	11	**
egetation	Broccoli	0.063		••	Leaves	60 days	***	.00	
egetation	Broccoli	0.171	**	**	Roots	60 days	**	**	**
egetation	Cauliflower	0.079		**	Leaves	70 days	•	11	**
egetation	Cauliflower	0.019	22.	**	Roots	70 days	**		11
egetation	Peas	0.001	n	**	Seeds	95 days	•	11	•
egetation	Peas	0.005	iii	**	Pods	95 days	11	11	•
egetation	Peas	0.110		11	Vines	95 days			
egetation	Peas	0.011	"	11	Roots				
getation	Radishes	Ø.237		ï	AND MEDICAL	95 days			
getation	Radishes	0.013			Tops	45 days			
,	Kadiones	0 . O T 3		_ (TE)	Tubers	45 days	**	**	***

27/1	,	Level	Hazard	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Medium	Use	means in ()	Response	Patnway	Receptor	Duracion	Hethou	beceing	KCICICHOC
Vegetation	Carrots	Ø.163	Background	Plant uptake	Roots	130 days	HNO3/HC104/ FLAAS	Greenhouse Soil pots	/ John (1972)
Vegetation	Beans	<0.1		**	Edible portions	Maturity "	FLAAS	Field"	Elfving et al (1978)
**	Cabbage	Ø.1		**	"	**			
Vegetation	Carrots	Ø.1			**		11		
Vegetation	Onions	0.2		. 11			**	11	**
Vegetation	Potatoes	Ø.1	**	***	11	**	11	11	
Vegetation		0.1			**		**	"	
Vegetation	Tomatoes Corn	Ø.027	**		Leaves	**		**	Chaney (1973)
Vegetation	Corn	0.0052	**		Grain	11	u ·	11	'n
Vegetation		0.0032	11	••	Leaves	**	1111	.11	n
Vegetation	Corn	0.002	11		Grain				11
Vegetation	Corn	0.062	11	H	Leaves		1	11	**
Vegetation	Soybeans	0.0028	11		Grain	11		**	
Vegetation	Soybeans	0.0020	11	H	014111	NR	NR	NR	Smart (1968)
Vegetation	Wheat/Barley	0.0053-0.012	11			Maturity	FLAAS	Field	Dudas and
Vegetation	Wheat	0.0053-0.0067	(well drained	11	**	"	"	"	Pawluk (1977)
S		a a) aa	(well drained	"	11	**			и (
Vegetation	Oats	0.0100	**	.11		11	**	**	
Vegetation	Barley	0.0060-0.0080	H.	"			**	**	
Vegetation	Wheat	0.0057-0.0063	(poorly drain						
		~ ~104	(poorly drain	iea "		**	**		,,
Vegetation	Oats	0.0120		11		**	#	11	"
Vegetation	Barley	0.0063-0.0067		11	**	NR	NR	NR K	abata-Pendias and
Vegetation	Barley (USA)	0.019		11	***	NR	NR NR	NR NR	Pendias (1984)
Vegetation	Oats (USA)	0.012		11	**	NR	NR NR	NR NR	" (1904)
Vegetation	Wheat (USA)	0.014				NR	NR	NR NR	
Vegetation	Wheat (USA)	Ø.010-0.016 WW			Soft Flo		NR NR	NR NR	Kitagishi an
Vegetation	Wheat (Japan)	Ø.02 WW							Yamane (1981
Vegetation	Oats	0.009	11	"	Grain	100 days	HNO3/HC104/	Greenhouse	/ John (1972)
Vegetation	Oats	0.107	**	**	Husks		FLAAS	Soil Pots	
Vegetation	Oats	0.176		"	Leaves	**		"	
Vegetation	Oats	0.011	**	**	Stalks	"	"		:
Vegetation	Oats	0.151	11	11	Roots	n			
Vegetation	Millet	0.1		"	Above ground biomass	Maturity	FLAAS	Field	Elfving et a (1978)

Table 8. Elevated mercury levels in plants.

Vegetation Leaf Lettuce 0.045 Not Noted Soil Solution Leaves 35 Days HNO ₃ /HO Vegetation " 0.387 " " 20 ugHg/g soil Roots 35 Days "	Study ethod Setting Reference ### Soil Pots ### ################################
Medium Use means in () Response Pathway Receptor Duration Medium Vegetation Leaf Lettuce 0.045 Not Noted Soil Solution Leaves 35 Days HNO ₃ /HO Vegetation " 0.387 " " 20 ugHg/g soil Roots 35 Days "	ethod Setting Reference HClO4/FLAAS Greenhouse/ John (1972) Soil Pots " " " "
Vegetation Leaf Lettuce 0.045 Not Noted Soil Solution Leaves 35 Days HNO ₃ /HC Vegetation " 0.387 " " 20 ugHg/g soil Roots 35 Days "	HC104/FLAAS Greenhouse/ John (1972) Soil Pots """"""""""""""""""""""""""""""""""""
Vegetation " 0.387 " " 20 ugHg/g soil Roots 35 Days	Soil Pots
Vegetation " 0.387 " " 20 ugHg/g soil Roots 35 Days	Soil Pots
	n
	n n n
Vegetation Spinach 0.695 " " Leaves 55 Days ' Vegetation " 1.067 " " Roots 55 Days '	n n n
Vegetation Broccoli 0.029 " " Leaves 60 Days '	W W
Vegetation " 1.870 " " Roots 60 Days "	
	и и и
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	II II II
	11 11 11
	II II II
Vegetation " 1.415 " " Roots 95 Days '	11 11 11
	и и и
	11 11 11
and the property of the control of t	11 11 11
vegetation 0.039	
	ry Analyzer Greenhouse/ Haney and
	" pot culture Lipsey (1973)
g e ga papara de la compania del compania del compania de la compania del la compania de la compania del la compania de la com	pot culture Lipsey (1973)
vector and the second s	
vegetation 3.4 ww 90.9% ik 2 Days	
vegetation 1.0 ww 88.9 ik	
vegetation 0.7 ww 68.88 ik	
vegetation 0.8 ww 11.0 YR " 10 Days	
vegetation 0.2 ww 0.08 ik 10 Days	XRFL Greenhouse/ Davis et al.
Vegetation Barley 2-5 (3) 10% YR HgCl ₂ Solution Leaves/ 5 leaf/ XI Shoots stage	KRFL Greenhouse/ Davis et al. sand culture (1978)
	HClO4/FLAAS Field/Soil Weaver et al.
Vegetation Bermuda grass 6.4 Toxic sand Leaves 6 weeks hnog/his	" pots (1984)
Vegetation Bermuda grass 0.2 Sig. wt. Westwood silt Leaves 6 weeks	n poes (1904)
reduction loam	
	HClo4/K2CrO7 Greenhouse/ Lindberg et
3,	AAS soil pots al. (1979)
Piomass Piomass	and soil pots al. (1979)
Vegetation Alfalfa 9.8 Uncertain Almaden soil Roots 16 weeks	11 11
Vegetation Oats 0.020 Not Noted Soil solution/ Grain 100 Days HNO3/HG	HClO ₄ /FLASS " John (1972)
Vegetation Oats 0.266 " " 20 ugHg/g soil Husks 100 Days	
	m m m
Vegetation Oats 0.026 " " " Stalks 100 Days	11 11
Vegetation Oats 0.426 " " Roots 100 Days	
	NR Solution Ishizuka and Culture Tanaka (1962)
	NR " "

prevent disorders such as white muscle disease. While excessive levels of selenium in forage are known to cause selenium poisoning or "alkali disease".

The factors affecting selenium availability and uptake by plants include the form of selenium in soil, soil type, soil pH, climate, presence of other elements and plant species (Whanger 1974). The inorganic phases of selenium occur as elemental selenium, as metal selenide, as a substitute in sulfides, as selenite and as selenate. Organic selenium occurs in soil as a result of partially decayed seleniferous vegetation. plant available forms are selenate and organic selenium (Gough, et al. 1979). The slightly mobile selenides and selenium sulfides dominate in acidic, poorly drained soils with high organic matter levels. Selenites, which are moderately available to plants, exist in well drained, neutral pH soils. Alkaline, well oxidized soils may contain appreciable levels of the soluble and readily available selenate form (Allaway 1968b, Lakin and Davidson 1967, Paasikallio 1981). The presence of other elements in the soil which are chemically similar to selenium, particularly sulfur, will result in the decrease in selenium uptake by the plant (Whanger 1974). Plant uptake of selenium is also dependent on the plant species involved. Most agricultural species accumulate only a few ppm while indicator species such as those in the genus Astragalus can accumulate up to 10,000 ppm (Rosenfeld and Beath 1964).

The following sections present selenium data for soils and plants reported in the reviewed literature.

2.3.1 Total selenium levels in soils

Selenium is found throughout the lithosphere at concentrations seldom exceeding 0.05 ppm (Kabata-Pendias and Pendias 1984). A world wide average for total selenium in surface soils is 0.40 ppm. A review of the literature suggests that the background level of total selenium in soils of the United States varies from 0.005 to 4.0 ppm. Tables 9 and 10 summarize the

Table 9. Background total selenium levels in soils.

		Level (ppm DW)	Hazard	Exposure				Study	
Medium	Use	means in ()	Response	Pathway	Receptor	Duration	Method		Reference
Soil, Colorado,									
Surface horizon	Cultivated and Uncultivated	0.1-1.4 (0.23)	Background	Plant uptake	NR	NR	XRFL	Field (168 samples	Connor and Shacklette (1975
Soil, Eastern U.S., B horizon	•	Ø.1-1.4 (Ø.39)	"	**	NR	NR	XRFL	Field (1000 samples)	•
Soil, Western U.S., B horizon	in.	Ø.1-4.3 (Ø.25)		III.	NR	NR	XRFL	Field (1000 samples)	и
Soils, Western U.S.		0.1-2.0	н	•	NR	NR	NR	Field	Swaine (1955)
Soils, Massachusetts	Vegetables	2.4-5.1 (3.5)	•		NR	NR	INAA/RNA	A Field	Laul et al. (1977
Soils, Wash., Surface	Vegetables	1.7			NR	NR	. 9 1	Field	» •
Soils, Ontario,									
Clay, pH 6.3	Agricultural	0.209	**	u.	NR	NR	NR	Field	Levesque (1974)
oam, pH 6.8	"	0.321			NR	NR	NR		
Clay, PH 6.7		0.395		"	NR	NR	NR	**	-
lay, pH 6.3	W	0.744		"	NR	NR	NR	**	
lay, pH 4.5		0.530			NR	NR	NR		
Clay, pH 5.2		0.460		···	NR	NR	NR		
oam, pH 7.0 oam, pH 7.2	11	0.450			NR	NR	NR		
Coam, pH 7.1		Ø.425 Ø.652	"	"	NR	NR	NR		÷ i
Loam, pH 6.0		Ø.652 Ø.197		u.	NR NR	NR NR	NR NR	n	
Soils, U.S.	NR	0.005-4.0	u		NR	NR	NR	Field K	abata-Pendias and Pendias (1984)
Soils, World-wid	e NR	0.005-4.0 (0.40) "		NR	NR	NR	300	•
Soils, Helena Valley, pH 8.0	NR	0.07			NR		cid diges AAS analy		EPA (1986)

Table 9. Background total selenium levels in soils, continued.

Medium	Use	Level (ppm DW) means in (Hazard) Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Soil, Sandy Ont., Canada	Agricultural	0.10-1.32 (Ø.27) Background		NR		2SO4/HNO3 gestion AAS	Field	Frank et al. (1979
Soil, Loam Ont., Canada	Agricultural	Ø.13-1.67 (Ø.38) Background	Plant uptake	NR		2SO4/HNO3 gestion AAS	Field	Frank et al. (1979
Soil, Clay Ont., Canada	#*************************************	Ø.16-1.43 (Ø.48) "	11	NR	NR		n	w ,
Soil, Organic Ont., Canada	W	0.10-0.75 (Ø.34) "	n ,	NR	NR	n	W	w
Soil, Muck	8								
Canada, Surface	Garden	1.3	II.	**	Vegetables	NR	IPAA	Field	Chattopadhyay and
0-7.5 cm	u u	1.22		u		NR	IPAA	**	Jervis (1974)
7.5-15 cm	**	Ø.81	11	11		NR	IPAA	**	
15-22.5 cm	N	0.62			**	NR	IPAA	11	n
22.5-30 cm	11	1.05	u			NR	IPAA	99	•
30-37.5 cm	**	Ø.91	W W	11	**	NR	IPAA	11	
37.5-45 cm	III	Ø.53	W	n	w	NR	IPAA	11	u
Cail Misseumi			× **						
Soil, Missouri,	Cultivation	Ø.2-1.5 (Ø.	45) "	iii	Corn	Maturity	XRFL	Field	Connor and
0-15 cm	Cultivation		43)	.11	Soybeans	Maturity	II L	11010	Shacklette (1975)
"	ü	0.1-1.4 (0. 0.1-1.5 (0.	21)		Pasture	Maturity	31	. •	BlackTette (1975)
Soil, Missouri,				8					e
Surface horizon	"	0.1-2.7 (0.	28) "	"	NR .	NR	" (3	Field 00 sample	es)
Soil, Missouri B horizon	Native	0.1-3.4 (0.	43) "	**	NŘ	NR	, 11	Field	ш.
Soils, Canada								¥	
Ø-15 cm, pH 5.9	Alfalfa	0.31	Background	Plant uptake	Plant tops		Flouro- metrically		Van Ryswyk et al. (1976)
65-87 cm, pH 7.	1 "	Ø.22	n	n	**	**	,,	"	
0-15 cm, pH 7.2		0.24	11	, W			**		
81-103 cm, pH 7	. 5 "	Ø.29	***	11	**	n n	311	н	11
Soils, Canada	NR	a a3_2 (a 2	26) Background	Plant uptake	NR	NR	Flouro-	Field	McKeague and

Table 9. Background total selenium levels in soils, continued.

		Level (ppm DW)	Hazard	Exposure		a g	a	Study	
Medium	Use	means in ()	Response	Pathway	Receptor	Duration	Method	Setting	Reference
Soil, Berrier Co.	MT								*
pH (6.6) Soil, Wayne Co. N	Orchard	<0.1-0.61 (0.095)	Background	Plant Uptake	Apples	Maturity	XRFL	Field	Shacklette (1980)
pH (5.5) Soil, Gloucester (•	<0.1 "	**	·	Apples	m .	w		w
NJ pH (5.5)	100	<0.1 "		ű "	Apples	**	11	u .	u
Soil, Yakima Co. (pH (6.6)		<0.1-0.34 (0.11)		n	Apples	n	m		**
Soil, Mesa Co., Co pH (7.8)	11	<0.1-0.4 (0.13)		u	Apples	11	u	ű.	u ·
Soil, Twin Falls (ID pH (8.2)		<0.1-(0.21)	**		Potatoes	**	. 11	u	

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Soil, Surface	NR	10	"Phytotoxical excessive"	ly NR	NR	NR	NR	NR	EL-Bassan and Tietjen (1977)
		5			•	•		n	Linzon (1978)
	***	10	11	u	**	**		<u>.</u>	Kabata-Pendias (1979)
n		10			***	**	* W	11	Kloke (1979)
Soils	Buckwheat	76.6	"Plants died"	64 ppm Se added to soil	NR	Maturity	Colorimetrically	Field plo	Martin (1936)
Soils	Buckwheat	10.5-39.6		8-32 ppm Se added to soil	m ×	u			н
Soils, Clay loam	Wheat	30	"Rapid yellow- ing and death			NR	NR	Greenhouse	Hurd-Karrer (1934)

literature pertaining to background and elevated levels of selenium in soils.

Few articles have reported phytotoxic levels of selenium in soils. Much of the concern associated with excess selenium stems from the toxicity of seleniferous plants to grazing animals. Seleniferous soils (>5.0 ppm total selenium) often support vegetation that is toxic to animals, however, these soils are generally not toxic to the plants growing naturally on them (NRC 1976). Kabata-Pendias and Pendias (1984) reviewed literature (not available to the present authors) that reported total soil selenium concentrations of 5 to 10 ppm as being phytotoxically excessive. Hurd-Karrer (1934) reported the death of wheat seedlings when soil selenium concentrations reached 30 ppm in greenhouse studies. The growth of buckwheat plants has been retarded at soil selenium levels of 10.5 to 39.6 ppm (Martin 1936). Death of these buckwheat plants occurred at a total soil selenium concentration of 76.6 ppm.

2.3.2 Selenium levels in plants

Rosenfeld and Beath (1964) proposed the classification of plants based on their ability to accumulate selenium and their potential toxicity to livestock. Group 1 plants were termed primary indicator or accumulator species which could absorb from 100 to 10,000 ppm. Most notable of this group were the Astragalus species. Group 2 plants were secondary selenium accumulators that rarely contained more than a few hundred ppm selenium. Most cultivated crops, grains and native grasses were classified as Group 3 plants. These species rarely accumulated more than 30 ppm total selenium. Tables 11 and 12 summarize background and elevated levels of selenium in plants reported in reviewed literature.

Vegetation containing greater than 2.0 ppm total selenium can be toxic to animals consuming it (NRC 1980). However, the same vegetation could contain selenium levels in great excess of this before experiencing phytotoxic symptoms.



An appreciable amount of literature exists on selenium levels in agricultural and range plants (non-accumulator species) and indicates that background concentrations usually range from Ø to 84 ppm (Table 11). While selenium is probably not essential for vegetative growth, the soluble forms of selenium are readily absorbed by plant roots (Kabata-Pendias and Pendias 1984). Because of the various solubilities and chemical forms of selenium, it is difficult to correlate the amount of total selenium in soils with the tissue concentration of plants.

Little documentation has been found concerning the determination of phytotoxic selenium levels in plant tissue. Martin (1936) reported growth reduction in buckwheat plants containing 35 to 124 ppm and death of plants containing 127 ppm selenium. A reduction in growth occurred in tomatoes with 191 ppm selenium (Yopp et al. 1974). Soltanpour and Workman (1980) concluded that 360 ppm selenium in the tops of alfalfa was responsible for very low yields while 1000 ppm was highly toxic. Selenium hazard levels for soils and plants are discussed in Section 3.3.

2.4 Silver Levels in Soils and Plants

Naturally occurring silver is found in minute quantities throughout the oceans, lithosphere, soils, plants and animals. Silver is similar to copper in its geochemical characteristics and exists as simple cations $(Ag^+, Ag^{2+}, Ag0^+)$ and complexed anions $[Ag0^-, Ag(S_20_3)_2^{3-}, Ag(S_04)_2^{3-}]$ (Kabata-Pendias and Pendias 1984). Silver is absorbed and complexed by organic matter and is apparently immobile at pH >4 (Kabata-Pendias and Pendias 1984). The availability of silver to plants is low due to the very low solubility of most of its compounds. Silver has not been proven to be essential for plant life (Vanselow 1965). The soluble fraction is extremely toxic, particularly to microorganisms and fish (Cooper and Jolly 1970). Silver, however, is relatively harmless to higher animals, including man. Silver data for soils and plants are presented in the following sections.

Table 11. Background selenium levels in plants.

		Level (ppm DW)	Hazard	Exposure					
Medium	Use	means in ()	Response	Pathway	Receptor	Duration	Method	Study Setting	Reference
Vegetation, Missouri	Corn	0.01-0.5 (0.06)	Background	Plant uptake	Grain		2-3 Diamino- naphthalene	Field	Connor and Shacklette (1975
**	Soybean	0.04-1.25 (0.11)			Seeds	Maturity			
<u>u</u> × 1	Buckbrush	0.02-0.08 (0.03)	m		NR	NR			
w	Cedar	0.01-0.04 (0.02)			NR	NR	*		
	Shagbark Hickory	0.02-0.04 (0.02)			NR	NR			
•	Post Oak	0.01-0.04 (0.02)	Ti .		NR	NR	•		₩
**	White Oak	0.01-0.04 (0.019) "	••	NR	NR			. (8)
н	Willow Oak	0.01-0.3 (0.032)	w i		NR	NR			•
11	Shortleaf Pine	0.02-0.2 (0.062)			NR	NR			
99		0.01-0.25 (0.02)	. II		NR	NR	. •		₩
**	Sweetgum	0.01-0.4 (0.065)			NR	NR	•		
Vegetation	Sagebrush	0.08-4.8 (0.42)		**	NR	NR			W
Vegetation, Washington	Cheatgrass	<0.03	Background	Plant uptake	Interior portions	NR	INAA and RNAA	Field	Laul et al. (1977)
Vegetation, Worldwide	Grasses	.00121	, ₂ , "	•	NR	NR	NR	NR	Kabata-Pendias and Pendias (1984
**	Clovers or alfalfa	.00588		W s	NR	NR	NR	NR .	•
*	Hay or fodder	.00287		~ n i	NR	NR	NR	NR	
egetation, U.S	. Grasses	.0104	***	•	NR	NR	NR	NR	10 X
	Clover or alfalfa	.0388	Background	Plant uptake	NR	NR	NR		bata-Pendias and Pendias (1984)

Table 11. Background selenium levels in plants, continued.

Medium	Use	(ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Vegetation,									
U.S.	Hay or fodder	.0336		•	NR	NR	NR	NR	
Vegetation,									
Northwest U.S	Rangelands	0.01-0.78	"	•	Plant top	s NR	Flourometrically	Field (94 samples)	Carter et al (1970)
н	Forage and	0.0-1.24	in .	H		NR		Field	
	Hay crops				N			(361 samples)	
Vegetation,					Military				
Western U.S.	Wheat	0.01-25.0	"	**	NR	NR	NR	Field (710 samples)	Rosenfeld and Beath (1964)
	Wheat	0.01-30.0			Grain	NR	NR	Field	•
								(176 samples)	
Vegetation, South Dakota	Native grass	a.a-84.a			NR	NR	NR	Field	
	The second secon					****	***	(294 samples)	
	n 6 Native Spec	ies <1.0	° n		Grazing	NR	AAS	Field Fletche	
Columbia		a Sen			stock			(294 samples)	Brink (1969)
Vegetation	Lettuce	0.002 WW	Background	Plant uptake	NR	NR	Acid digestion	NR	Wolnik et al. (1983)
	Peanuts	0.057 WW	**	•	NR	NR	FLAAS analysis	NR	(1505)
	Potatoes	0.003 WW	**	10	NR	NR	199	NR	•
	Soybeans	Ø.19 WW	11		NR	NR	•	NR	
	Sweet Corn	0.006 WW	11	11	NR	NR	•	NR	•
	Wheat	0.37 WW		•	NR	NR		NR	•
Vegetation,		a aar a aaa	"	n					
Canada	Timothy	0.005-0.023	"	,,	NR	NR	NR	Field	Gupta and
11	Red clover	0.004-0.031	"	11	NR	NR	NR	Field	Winter (1975)
	Oats	0.004-0.043			Kernal	NR	NR	Field	
	Barley	0.006-0.040			Kernal	NR	NR	Field	
Vegetation, Massachusetts	Corn	<0.03	Background	Plant uptake	Interior	NR	INAA and RNAA	Field	Laul et al.
massachusetts	Potatoes	<0.03	Background	riant uptake	portions	NR NR	INAA and KNAA	Liela	
"	Peas	<0.03	n		borcious	NR NR			(1977)
ce.	reas	\U.U.			20	NK	1(5,5)		**

Table 12. Elevated selenium levels in plants.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration		Study Setting	Reference
Vegetation, Illinois	Wheat	380	"No injury to plant"	NR	NR	NR	NR	NR	Yopp et al. (1974)
Vegetation, Illinois	Tomato	191	"Growth reduc- tion"	- NR	NR	NR	NR	NR	**
Vegetation	Buckwheat	127	"Plants died"	64 ppm Se added to soil	l NR	Maturity	Colorimetrically	NR	Martin (1936)
Vegetation	Buckwheat	35-124	"Growth retarded"	8-32 ppm Se added to soil	NR		п	NR	
Vegetation	NR	5-30	"Excessive or toxic"	NR	Leaf tissue		NR	NR Ka	bata-Pendias and Pendias (1984)
Vegetation .	Alfalfa	360	"Produced very low yields"	, NR	Plant to	p NR	Hot water extract	Greenhouse	Soltanpour and Workman (1980)
Vegetation	Alfalfa	1000	"Highly toxic"	' NR		NR	11	W	
Central Oregon	Alfalfa	0.13-0.34	No effect	Na ₂ SeO ₃ added to soil	Plant to	op 1 year	Allaway and Carey (1964)	Field	Allaway et al. (1966)



2.4.1 Total silver levels in soils

Silver is an element found universally in soils (Vanselow 1965). Literature reviewed by Smith and Carson (1977b) shows that background levels of silver in soils range from 0.1 to 5.0 ppm. Reported background silver levels in the United States indicate total silver concentrations in soils seldom exceed 0.5 ppm (Connor and Shacklette 1975). No literature has been found on extractable levels of silver in undisturbed soils. Tables 13 and 14 summarize the background and elevated silver levels in soils found in the reviewed literature.

Few studies have determined phytotoxic levels of silver in soils. Concern regarding silver pollution of soils has emerged recently due to increased use of silver iodide as a nucleating agent for promoting precipitation (cloud seeding). Research has shown that typical aerial fallout levels of silver from cloud seeding $(10^{-7} \text{ to } 3 \text{ x } 10^{-7} \text{ ppm})$ poses no immediate threat to the soil resource (Cooper and Jolly 1970). Aerial deposition of silver near a silver mine and treatment plant in New Zealand resulted in average soil silver concentrations (1.7 ppm) being significantly greater than background (0.2 ppm) concentrations (Ward et al. 1977). These elevated silver levels decreased with distance from the treatment plant. The only soil phytotoxic criteria found in the reviewed literature was that of Linzon (1978) who reported 2 ppm total soil silver was phytotoxically excessive.

2.4.2 Silver levels in plants

A study of 35 plant species, representing the major vascular plant groups, revealed background plant tissue silver concentrations ranged from 0.01 to 16.0 ppm (Horovitz et al. 1974). These authors noted higher silver content in some fungiand bryophytes and lower values in angiosperms and gymnosperms.

A large amount of literature published on silver concentrations in plants indicate that background concentrations usually range from Ø to 1.0 ppm (Table 15). Shacklette (1980) reported less than 1 ppm (ash weight basis) for most fruits and vegetables

0100700

Table 13. Background total silver revels in soils.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duratio	n Method	Study Setting	Reference
Soil, Missouri (0-15 cm)	Cultivated	<0.5-3 (<0.5)	Background	Plant uptake	NR	NR	6-step ES (1400	Field Samples) Sha	Connor and cklette (1975)
Soil, Missouri B-horizon	Oak-Hickory Forest	<0.5-3 (<0.5)	300.	n	NR	NR	n	Field (300 samples)	. n
Soil, Colorado (0-15 cm)	Cultivated and Uncultivated	<0.5-1.5 (<0.5)	W	u	NR	NR	H	Field (168 samples)	n
Soil, Western U.S. (20 cm)	Native Vegetation	<0.5-5 (<0.5)	и .	in:	NR	NR	n	Field (1000 samples	# 5)
Soils, Ontario Canada (0-15 cm)) Croplands	0.04-1.81 (0.44	1) "	in .	NR	NR	HNO ₃ Digestion AAS analysis	Field (228 samples)	Frank et al. (1979)
Soils, Worldwide	e Cultivated ar Native	nd <0.01-5.0	u	m	NR	NR	Spectrographically	Field	Swaine (1955)
Soils, Helena Valley pH 8.0	in:	0.25	u	"	NR	NR	Acid digestion, AAS analysis	Field	EPA (1986)
Soils, Surface Muck, Canada 0-7.5 cm 7.5-15 cm	Garden "	Ø.89 Ø.68 Ø.52	n u	" "	Vegetable " "	NR NR	IPAA "	Field and	Chattopadhyay d Jervis (1974)
15-22.5 cm	300	0.40		"		NR	NR	Field	Vanselow (1965)
Soils, Scotland	NR	<2.0	"		NR	NR		Field	"
Soils, Missouri	Agricultural	Ø.7	. "		NR	NR	NR		n
Soils, Californ	ia NR	0.2-0.7	" .		NR	NR	NR	Field	
Soils, surface New Zealand	Native	0.21		**	NR	NR	Acid disgestion,	Field	Ward et al. (1977)
Aberdeenshire U	K NR	0.29-0.50		•	NR	NR	SSMS	Field	Ure and Bacor (1978)
Tubingen Univ Germany	Botanical Garden	0.08-0.09	11	" Ephe	Juniperus Communis- dragerardia	na NR	RNAA	Field	Horovitz et al. (1974)

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Soils, surface	NR -	2.0	"Phytotoxically excessive"	Plant uptak	e NR	NR	NR	NR	Linzon (1978)
Soils, surface New Zealand	Native	0.75-3.3(1.7)	"Significantly higher than background"	Aerial fallout	NR	NR	Acid digestion, AAS analysis	Field	Ward et al. (1977)

Table 15. Background silver levels in plants.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Vegetation	Vascular plants	5.0	Background	Uptake from soil	Plant	NR	NR	NR	Shacklette (1965)
Vegetation	Gymnosperms	0.07	w	w	n	NR	NR	NR	Bowen (1966)
Vegetation British Columbi	ia Gymnosperms	0-1.4	m	н	n	NR	Fire assay	Field	Warren and Delavault (1956
Vegetation, U.S	S. Angiosperms	0.06			Plant to	ps NR	ES		Cannon et al. (1968
Vegetation British Columbi	ia Angiosperms	Ø28		II.	Plant	NR	Fire assay		Warren and Delavault (1959
/egetation	Grains and cereals	0.9		n	***	NR	NR	NR	Browning (1961
	Generalized	Ø.5	in .	n,	Leaf tissue	Maturity	NR	NR	Kabata-Pendias an Pendias (1984)
egetation, eorgia	Snap Bean	<0.5	n	**	Edible portions	Maturity	Plant ash, 6-step ES	Field	Connor and Shacklette(1975
egetation	Cabbage	<0.5	***	II .	u	Maturity	"	n	z 10
n	Tomato	<0.5	, "	n	n n	Maturity	H		
in.	Alfalfa	0.1-0.5	n	**	Tops	NR	NR	Field	Vanselow (1965
w	Bur clover	0.2-0.5			n	NR	NR	**	
	Ladino clover	0.4-0.6	90	11	W	NR	NR	m .	11
m .	Grasses	0.1-0.4				NR	NR	11	,
**	Wheat	Ø.4	ii .	u	Whole grain	Maturity	NR	10	n
egetation Powd idge Basin	der Big Sagebrush	<1.0	Background Geometric mea	" an	Plant	NR	Plant ash, ES		Connor et al (1976)

		Leve (ppm	DW)	Hazard	Exposure				Study	Reference
Medium	Use	means	in ()	Response	Pathway	Receptor	Duration	Method	Setting	Reference
Vegetation, New Zealand	Tawa (tree)	Ø.22		Background	Uptake from	Washed leaves	NR	Ashed, AAS	Field	Ward et al. (1977)
n	Ü	Ø.24		n	"	Unwashed leaves	NR		11	
u	W	0.20		u	u ,	Washed twigs	NR	u	•	•
	Perennial ryegrass	0.06		n	"	Roots	NR		**	91
n	11	0.08		•	u	Leaves	NR	w	•	, n
	White clover	0.08		и	111	Roots	NR		и "	n
,,	W	Ø.1Ø			w	Leaves	NR		**	и
•	Annual bluegrass	Ø.Ø6		11		Roots	NR	"H	**	,,
* 1 11	**	Ø.07			**	Leaves	NR		11	w
	Cocksfoot	Ø.1Ø		'n		Roots	NR	n	11	,,
· (m)	S	0.10		11		Leaves	NR		11	w.
,,	Yorkshire fog	0.06		"		Roots	NR	u		11.
w		Ø.Ø8			n	Leaves	NR			n
*	Flatweeds	Ø.12			w	Roots	NR	11		n
11	"	Ø.14		W	"	Leaves	NR	n		'n
n .	Birdsfoot treefoil	Ø.Ø8			u	Roots	NR	,11		, 10
	·ii	Ø.Ø8			•	Leaves	NR	···	n	311

Table 15. Background silver levels in plants, continued.

Medium	Use	Level (ppm DW) means in ()	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Commercial Farms	Vineyard	<0.045	Background	Plant Uptake	American Grapes				
USA	vineyala	(0.045	background	riant opeane	- Fruit	NR	ES	Field	Shacklette (1980)
(1980)			**		31 B	NR	ES	Field	(1300)
" "	Orchard	<0.016-0.032	u	u	Apples-Fruit	NK	ES	rieid	
••	Vineyard	<0.027			European Grapes	NR	ES	Field	
_			**		- Fruit			Field	
	Orchard	<0.067-0.134			Peaches-Fruit	NR	ES	Field	•
	"	<0.021	111		Pears-Fruit	NR	ES		
"		<0.048-0.144			Plums-Fruit	NR	ES	Field	=
11	Vegetables	<0.093	**	••	Cabbage-Heads	NR	ES	Field	
"	11	<0.071	**	311	Carrots-Roots	NR	ES	Field	
•		<0.100	**	••	Cucumbers-Fruit	NR	ES	Field	14 To 100
••	11	<0.039-0.117	**	**	Dry Beans	NR	ES	Field	
. 11	n	<0.140-0.28	n	**	Lettuce-Heads	NR	ES	Field	
**	11	<0.042-0.042	**	88	Potatoes-Tubers	NR	ES	Field	**
••	11	<0.070		••	Snap Beans-Pops	NR	ES	Field	
	n	<0.026	11	•••	Sweet Corn-Grains	NR	ES	Field	•
11	W	<0.120	n	" m	Tomatoes-Fruit	NR	ES	Field	•
m .		<0.100	••		Asparagus-Shoots	NR	ES	Field	*
11	**	<0.098	**	**	Cantaloupes-Fruit		ES	Field	
**	n	<0.200	"	,	Chinese Cabbage				
					- Leaves	NR	ES	Field	•
	11	<0.074	iii	**	Eggplant-Fruit	NR	ES	Field	•
		<0.074	**	••	Endive-Leaves	NR	ES	Field	•
***					Onions-Bulbs	NR	ES	Field.	
**	···	<0.042	W	· · · · · · · · · · · · · · · · · · ·	Parsley-Leaves	NR NR	ES	Field	
		<0.190	•			MK	100	. 1010	
-	***	<0.084	••		Fresh Peppers	NR	ES	Field	
					- Fruit	NK	E0	Fierd	

tested (25 species). The maximum silver concentration was found in fresh lettuce (0.28 ppm DW). The amount of silver taken up by plants is related to the amount of the metal in the soil (Kabata-Pendias and Pendias 1984). Therefore, elevated levels of silver could accumulate in plants growing in soils enriched with silver due to aerial deposition from smelting. Tables 15 and 16 present background and elevated data for silver levels in plants.

Little research has been conducted on the determination of excessive levels of silver in plant tissue. Ward et al. (1977) reported that a mean of 1.2 ppm in roots and 1.8 ppm in leaves of pasture species were significantly higher than background levels. A 10% yield reduction occurred in spring barley with 4 ppm silver in plant tops (Davis et al. 1978). Bush bean yields were reduced with 1760 ppm silver in roots, 5.1 ppm in stems and 5.8 ppm in plant tops (Wallace et al. 1977b). Silver hazard levels for plants and soils are discussed in Section 3.4.

2.5 Thallium Levels in Soils and Plants

Thallium is a rare element that is found in trace quantities in most soils and geological materials. Thallium exists in both mono $(T1^{+1})$ and trivalent $(T1^{+3})$ states and compounds of both forms are highly toxic (Logan et al. 1983). Monovalent thallium forms "sparingly" soluble compounds similar to the heavy metals copper, silver, gold, mercury and lead. The anion of these compounds include sulfides, iodides, chlorides and chromates (Smith and Carson 1977a). Trivalent thallium is usually found only in very acid environments (Smith and Carson 1977a). Thallium is geochemically similar to the alkali metals (potassium, rubidium and cesium) and is found as an isomorphic substitution in potassium feldspars (orthoclase, microcline, sanidine), micas and potassium feldspathoids (leucite) (Wedephol 1978). The element is also found in many metalic sulfide ores including sphalerite, pyrite and galena (Smith and Carson 1977a). is usually disseminated in low temperature hydrothermal deposits of antimony, mercury, lead and zinc and ores high in arsenic content have been found to be enriched in thallium (Velikii et

Table 16. Elevated silver levels in plants.

Medium	Use	Level (ppm DW) means in	Hazard () Response	Exposure Pathway	Recentor	Duration	Method	Study Setting	Reference
			, Kesponse	ruchway	Receptor	Durucion	Heemod	beccing	Kererence
Vegetation	Bush Bean	1.0	"No toxicity"	Nutrient Solu- tion Ø ppm Ag	- Roots	13 Days	Emission Spectrography	Greenhous	e Wallace et al. (1977b)
Vegetation	н	Ø.3	"	**	Stems	•	m	•	
Vegetation	•	0.2			Leaves	m	n	m	•
Vegetation	"	83	Ü	Nutrient Solu- tion 0.108 ppr Ag		"	**	н	W
Vegetation	и	0.8		**	Stems	••	**		
Vegetation		1.0	"	**	Leaves		**	•	
Vegetation	•	1760	"Yields greatly decreased"	Nutient Solu- tion 1.08 ppm Ag	Roots	**	**	n	
Vegetation	'n	5.1			Stems		n	*	
Vegetation		5.8	"	•	Leaves	•	, "	•	•
Vegetation	Spring Barley	4.0	"10% yield reduction"	Sand Culture I Nutrient Solu- tion		27 Days	Tri-acid digest Colorimetric an		Davis et al. (1978)
Vegetation New Zealand	Pasture species	1.2	"Significantly higher than background"	y Plant uptake Aerial fallo		NR	Ashed, Atomic Absorption	Field	Ward et al. (1977)
Vegetation New Zealand	•	1.8		n	Leaves	NR	**		
Vegetation	NR	5-10	"Excessive or Toxic"	NR	Leaf tissue	Maturity	NR	NR	Kabata-Pendias an Pendias (1984)

al. 1968). Many gold ores are commonly enriched in thallium (Zimmerley 1947). Thallium is commonly present in coal at approximately 0.7 ppm, probably as sulfide inclusions (Smith and Carson 1977a).

Thallium has been used in the past as an insecticide and rodenticide but has been banned from these products used in the United States since 1972 (Smith and Carson 1977a). Carlson et al. (1975) have reported thallium salts were most phytotoxic of thallium, lead, cadmium and nickel salts that were tested on hydroponically grown corn and sunflowers.

Thallium is released to the environment from combustion of coal and from smelting operations. It is used primarily in electrical component manufacturing (Smith and Carson 1977a). An assessment of anthropogenic deposition of thallium suggested little or no increase over present levels is expected in the future (Galloway et al. 1982), but local areas may be impacted by thallium pollution (Scholl and Metzger 1981).

2.5.1 Total thallium levels in soils

Few reports have been published on the characteristics of thallium in soils. Thallium is easily mobilized and transported together with alkaline metals (Kabata-Pendias and Pendias 1984) and is apparently readily available to plants (Scholl and Metzger 1981, Hoffman et al. 1982). Thallium is immobilized in soils through fixation by clays and manganese or iron oxides, and can be sorbed by organic matter (Kabata-Pendias and Pendias 1984). Thallium may also be removed from the soil solution by base exchange (McCool 1933).

Thallium occurs in trace amounts in most rocks but is found in higher concentrations in acid rocks (granites, gneisses) than in mafic or ultra mafic rocks (basalts, gabbros, dunites, peridotites and pyroxenites) (Bohmer and Pille 1977, Kabata-Pendias and Pendias 1984, Smith and Carson 1977a). Background levels of thallium in igneous rocks range from 0.05 to 2.3 ppm. Thallium is found at higher concentrations in fine grained (claystone/shale) sedimentary rocks as compared to coarse grained

rocks. Typical thallium levels in shales and sandstones have been reported as 0.5 to 2.0 ppm and 0.4 to 1.0 ppm respectively (Kabata-Pendias and Pendias 1984). Bowen (1966) gave a typical background soil level of 0.1 ppm (Table 17). Background thallium values for surface muck (gley) soils have been reported at 0.20 and 0.22 ppm (Chattopadhyay and Jervis 1974). Samples of the muck soil from 0 to 7.5 cm and 22.5 to 30 cm depths exhibited thallium levels of 0.17 and 0.18 ppm respectively.

Little data are available on the effect of elevated thallium levels in soils and the resulting effect to plant production (Table 18). McCool (1933) has reported the injury to corn, ryegrass and wheat was not reduced by leaching soil with up to 91.5 cm (36 in) of water, but the extremely high concentrations of the Tl₂SO₄ (0.02 ml Tl₂SO₄/q soil) made these data of little use. Solution culture experiments by Potsch and Austenfeld (1985) and Pieper and Austenfeld (1985) have indicated concentrations of 10uM T1NO3 or 10uM T1(NO3)3 (10 uM = 2 ppm) produced significant reductions in the dry matter yields of pea plants but not in faba beans. Thallium levels up to 4.5 ppm have been reported in soils near an abandoned cement kiln plant (Scholl and Metzger 1981). These authors documented increased plant uptake of thallium from the polluted soils and noted thallium specific toxicity symptoms in some plants, but did not determine the effect on yield.

2.5.2 Thallium levels in plants

Few studies have investigated the toxicity of thallium to higher plants. The metal has not generally benefited from the large mass of data generated by sewage sludge disposal problems. Experimental data suggest that an increase in soil thallium levels increases uptake by plants (Hoffman et al. 1982, Scholl and Metzger 1981).

Experiments with barley roots suggested monovalent thallium was absorbed at a steady rate while trivalent thallium reached a plateau level in a short time (30 minutes in solution culture) (Logan et al. 1983). These authors found trivalent thallium was

Medium	Use	Level (ppm DW)	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Essen Soil Dautmergen	ND Crops	Ø.1 3.0	Background Background	Plant Uptake Plant Uptake	Vegetables Vegetables	ND Maturity	Photometric Photometric	Field Greenhou se	Scholl and Metzger (1981) Hoffmann et al. (1982)
Ritzville Silt Loam Scottland Topsoil Scottland Topsoil	Soybeans NR NR	0.33 0.17 0.37	Background Background Background	Plant Uptake NR NR	Leaves, Stem Pods NR NR	60 days NR NR	AAS SSMS SSMS	Soil Pots Field Field	Cataldo and Wildung (1978) Ure and Bacon (1978) Ure and Bacon (1978)
Canadian Muck Soil Canadian Muck Soil Canadian Muck Soil		<pre>@ .0.21 (SUR) @ .17 (0-7.5 cm) @ (22.5-30.0 cm)</pre>		Plant Uptake Plant Uptake Plant Uptake	Garden Vegetables	NR NR NR	IPAA IPAA IPAA	Field Field Field	Chattopadhyay and Jervis (1974) Chattopadhyay and Jervis (1974) Chattopadhyay and Jervis (1974)

Table 13. Elevated total thallium levels in soils.

	The second secon	Level	Hazard	Exposure			W-41-4	Study Setting	Reference
Medium	Use	(ppm DW)	Response	Pathway	Receptor	Duration	Method	secting	Reference
Dautmergen tL	Kohlrabi	503	24 % YR	Plant Uptake					
ouurmer yen to	Nozub.	303		(TINO3)	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
	Radish	503	51 % YR		Tubers				
**	Radish	503	No YR		Leaves				
	Green Rapeseed	503	91 % YR		Tops	•	•	•	•
	Lettuce	503	73 % YR		Leaves	•		•	•
Dautmergen tL	Kohlrabi	203	1 % Yield	Plant Uptake					
150			Increase	(T1NO3)	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
	Radish	203	21 % YR	•	Tubers		I	-	<u> </u>
	Radish	203	8.7 % YR	•	Leaves			=	
	Green Rapeseed	203	38 % YR	•	Tops	•		-	
	Lettuce	203	62 % YR	•	Leaves	•		-	•
Dautmergen tL	Kohlrabi	53	12 % Yield	Plant Uptake					
			Increase	(T1NO3)	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
•	Radish	53	39 % YR	"	Tubers	•			
	Radish	53	11 % YR		Leaves	•	-	-	
	Green Rapeseed	53	2.9 % YR		Tops				
	Lettuce	53	44 % YR		Leaves		₽■0		•
Dautmergen tL	Kohlrabi	13	12 % Yield Increase	Plant Uptake (T1NO3)	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffman et al. (1982)
	Radish	13	10 % Yield					19 <u>24</u> 1	
			Increase	**	Tubers		•	1	
	Radish	13	17 % YR		Leaves			11 1	
	Green Rapeseed	13	2.9 % YR		Tops		•	•	
	Lettuce	13	23 % YR	•	Leaves	**	•	•	

readily desorbed by plants compared to monovalent thallium and concluded monovalent thallium was absorbed by plants in competition with potassium and therefore dependent on metabolic energy, whereas trivalent thallium was not. Monovalent thallium is apparently the most readily accumulated by plants due to its ionic radius which is similar to potassium and the element thus mimics potassium in many biological processes (Logan et al. 1983). Cataldo and Wildung (1978) demonstrated a 57 percent reduction of thallium uptake in the presence of a 10 fold increase in the potassium concentration. Thallium partitioning in plant parts is apparently very species specific. Work by Potsch and Austenfeld (1985) and Pieper and Austenfeld (1985) indicated that pea plants (Pisum sativum L.) concentrate thallium (as $T1NO_3$, $T1(NO_3)_3$, $T1^{+1}$ EDTA and $T1^{+3}$ EDTA) in stems while field beans (Vicia faba L.) concentrate thallium in roots. authors found thallium levels in pea leaves to be consistently higher than thallium levels in bean leaves in plants grown in the same concentration of thallium.

Background data for thallium levels in vegetation has been reported by several authors, including Geilmann et al. (1960) and Schacklette et al. (1978) (Table 19). Levels of thallium in plant tissues are generally much less than 1 ppm (Smith and Carson 1977a). However, values range from 0.008 ppm in clover to 35 ppm in kohlrabi.

Uptake of thallium by plants exposed to elevated thallium levels in soils follows the plant specific pattern. Green cabbage, which exhibits relatively higher background thallium levels (Geilman et al. 1961 and Hoffmann et al. 1982) also accumulates higher amounts under elevated conditions (Hoffman et al. 1982). Turnip leaves and rape plants can accumulate high levels of thallium (Hoffman et al. 1982). Scholl and Metzger (1981) demonstrated rape plants uptake 5 to 8 percent of applied thallium and 2 to 5 percent of natural soil thallium, and suggested the use of rape plants to decontaminate thallium polluted soils. These authors also noted green kale, turnips,

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Level Hazard Exposure Study Medium Use (ppm DW) Response Pathway Receptor Duration Method Setting Reference Vegetation Subalpine Fir 2-100 AWT Background Needles Rocky Mountains Shacklette et al. (1978) Vegetation Subalpine Fir 2-70 AWT Background Minimal NR NR Stems Rocky Mountains Shacklette et al. (1978) Vegetation Limber Pine 2-5 AWT Background Minimal Needles NR NR Rocky Mountains Shacklette et al. (1978) Vegetation Rocky Mountains Limber Pine 3-5 AWT Background Minimal Stems NR NR Shacklette et al. (1978) Vegetation Lodgepole Pine 2-5 AWT Background Minimal Needles NR Rocky Mountains Shacklette et al. (1978) NR Vegetation Lodgepole Pine 3-7 AWT Background Minimal NR Rocky Mountains Stems NR Shacklette et al. (1978) Vegetation Engelmann Spruce 2-10 AWT Background Minimal NR NR Needles Rocky Mountains Shacklette et al. (1978) Vegetation Engelmann Spruce 15 AWT Background Minimal Stems NR NR Rocky Mountains Shacklette et al. (1978) Vegetation Myrtle Blueberry 2-7 AWT Background Plant Uptake Stems/Leaves NR NR Rocky Mountains Shacklette et al. (1978) Vegetation Ponderosa Pine 15 AWT Background Minimal Stems NR Rocky Mountains Shacklette et al. (1978) Vegetation Clover 9.998-9.919 Background Plant Uptake NR NR NR Field Geilmann et al. (1961) Vegetation Meadow Hay 0.02-0.025 Background Plant Uptake NR NR Field Geilmann et al. (1961) Vegetation Head Lettuce 0.021 Background Plant Uptake NR NR NR Field Geilmann et al. (1961) Vegetation Red Cabbage 0.040 Background Plant Uptake NR Geilmann et al. (1961) Field Vegetation Green Cabbage 0.125 Background Plant Uptake NR NR NR Field Geilmann et al. (1961) Vegetation Leek 0.075 Background Plant Uptake NR NR NR Field Geilmann et al. (1961) Vegetation Endive 0.080 Background Plant Uptake NR NR Geilmann et al. (1961) Field Vegetation 0.025-0.030 Reet Background Plant Uptake Leaves NR Geilmann et al. (1961) Field Vegetation Potato 0.025-0.030 Background Plant Uptake Above Ground Biomass NR Field Geilmann et al. (1961) Vegetation Kohlrabi 3.7 ppm Tl Plant Uptake Leaves NR Photometric Hoffmann et al. (1982) ND Vegetation 3.7 ppm T1* Kohlrabi 0.10 Plant Uptake Tubers NR Photometric ND Hoffmann et al. (1982) Vegetation Zucchini 0.90 5.2 ppm T1* Plant Uptake NR Photometric ND Leaves Hoffmann et al. (1982) Vegetation Zucchini 5.2 ppm Tl* Plant Uptake NR Photometric ND Stems Hoffmann et al. (1982) Vegetation Cucumbers 0.70 5.4 ppm T1* Plant Uptake Leaves NR Photometric ND Hoffmann et al. (1982) Vegetation Cucumbers 0.10 5.4 ppm T1* Plant Uptake Fruit NR Photometric ND Hoffmann et al. (1982) Vegetation Red Beet 2.40 5.2 ppm T1* NR Plant Uptake Leaves Photometric ND Hoffmann et al. (1982) Vegetation Red Beet 5.2 ppm T1* 0.60 Plant Uptake Tubers NR Photometric ND Hoffmann et al. (1982) Vegetation Carrots 0.30 3.5 ppm T1* Plant Uptake Leaves NR Photometric Hoffmann et al. (1982) Vegetation 3.5 ppm T1* Carrots 0.10 Plant Uptake Roots Nr Photometric ND Hoffmann et al. (1982) Vegetation 0.9 ppm T1* Onions 0.10 Plant Uptake Tops Photometric Hoffmann et al. (1982) Vegetation 0.9 ppm T1* Onions 0.01 Plant Uptake NR Photometric Hoffmann et al. (1982) Tubers Vegetation 3.0 ppm T1* Kohlrabi 30.0 Plant Uptake Old Leaves Maturity Photometric Greenhouse/Soil Pots Hoffmann et al. (1982) Vegetation 3.0 ppm T1* Kohlrabi 6.0 Plant Uptake Young Leaves Maturity Photometric Greenhouse/Soil Pots Hoffmann et al. (1982) Vegetation 3.0 ppm T1* Green Rapeseed 10.0 Plant Uptake Maturity Photometric Greenhouse/Soil Pots Hoffmann et al. (1982) Tops 3.0 ppm T1* Vegetation Radish 1.1 Plant Uptake Tubers Maturity Photometric Greenhouse/Soil Pots Hoffmann et al. (1982) 3.0 ppm T1*
3.0 ppm T1* Vegetation Radish Plant Uptake Leaves Maturity Photometric Greenhouse/Soil Pots Hoffmann et al. (1982) Vegetation Lettuce Plant Uptake Hoffmann et al. (1982) Leaves Maturity Photometric Greenhouse/Soil Pots

Table 19. Background thallium levels in plants.

^{*} Soil

Table 20. Elevated thallium levels in plants.

No.	1900	Level	Hazard	Exposure			507 899	Study	P44
Medium	Use	(ppm DW)	Response	Pathway	Receptor	Duration	Method	Setting	Reference
D	C D	3326	91 % YR	m1 vo			Dh		
Dautmergen Soil Dautmergen Soil	Kohlrabi	2354	21 % YR	T1NO3	Tops Old Leaves	Maturity Maturity	Photometric Photometric	Greenhouse/Soil Pots Greenhouse/Soil Pots	Noffmann et al. (1982) Noffmann et al. (1982)
Dautmergen Soil	Kohlrabi	1936	9.3 % YR	TINO3	Old Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Moffmann et al. (1982)
Dautmergen Soil		1656	38 % YR	TINO3			Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Dautmergen Soil	Kohlrabi	1080	25 % YR	TINO3	Tops Young Leaves	Maturity Maturity	Photometric	Greenhouse/Soil Pots	Moffmann et al. (1982)
Dautmergen Soil	Kohlrabi	1011	7 % YR	TINO3	Old Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Dautmergen Soil	Kohlrabi		Yield Increase		Young Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Dautmergen Soil		499	2.9 % YR	TINO3	Tops	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	Pea	440	56 % YR	Tl(I) EDTA	Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Dautmergen Soil	Kohlrabi	382	7 % YR	TINO3	Old Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	Pea	360	59 % YR	TINO3	Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Dautmergen Soil	Radish	331	No Yr	TINO3	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	Pea	320	47 % YR	Tl(III) EDTA		ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Dautmergen Soil			% Yield Increase		Young Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	Pea	233	46 % YR	Tl(I) EDTA	Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Sand Culture	Pea	230	41 % YR	T1(III) EDTA		ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Field Bean	222	18 % YR (N.S.)		Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Solution Cultur	and the constitute and the constitution	210	32 % YR (N.S.)		Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Solution Cultur			Yield Increase			ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Field Bean		% Yield Increase					oreeouse	THE PART OF THE PARTY OF THE PARTY
			(N.S.)	Tl(I) EDTA	Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Dautmergen Soil	Green Rapeseed	180	2.9 % YR	TINO3	Tops	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Dautmergen Soil	Radish	150	8.7 % YR	TINO3	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	Pea	140	70 % YR	Tl(I) EDTA	Leaf	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Sand Culture	Field Bean	130	5.1 % YR	Tl(III) EDTA		ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Pea	123	6.3 % YR (N.S.)		Stem	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Pea	120	69 % YR	Tl (III) EDTA	Leaf	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Dautmergen Soil	Kholrabi	116 24 9	% Yield Increase	T1NO3	Young Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	Pea	115	25 % YR	T1 (NO3) 3	Stem	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Pea	110	60 % YR	Tl(I) EDTA	Leaf	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Sand Culture	Field Bean	108	11 % YR (N.S.)	TlNO3	Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Sand Culture	Field Bean	103	51 % YR	Tl(III) EDTA	Leaf	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Field Bean	88 15	% Yield Increase						
			(N.S.)	Tl(I) EDTA	Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Sand Culture	Pea	86	61 % YR	T1NO3	Leaf	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Sand Culture	Corn	82	50 % Reduction						
No. M. She had to		Market	Photosynthesis		Leaf	4-5 days	AAS	Hydrophonic/Greenhouse	
Sand Culture	Field Bean	76	11 % YR	$T1(NO_3)_3$	Stem	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Pea	75	45 % YR	T1NO3	Leaf	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Dautmergen Soil		64.4	11 % YR	T1NO3	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	Pea	63	6.7 % YR (N.S.)		Stem	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Solution Cultur	e Sunflower	63	50 % Reduction					8	
	160	tanan	Photosynthesis	AND THE PARTY OF T	Leaf	4-5 days	AAS	Hydrophonic/Greenhouse	
Sand Culture	Pea	58	30 % YR	Tl(I) EDTA	Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Sand Culture	Pea	43	8 % YR (N.S.)		Leaf	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Field Bean	36	47 % YR	T1 (NO3) 3	Stem	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Pea	36	17 % YR (N.S.)		Stem	ND	AAS	Greenhouse	Pieper and Austenfeld (1985)
Sand Culture	Pea	36	29 % YR (N.S.)		Stem	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Dautmergen Soil		35.1	21 % YR	TINO3	Tubers	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	Pea	35	31 % YR	Tl(III) EDTA	Leaf	ND Material Land	AAS	Greenhouse	Pieper and Austenfeld (1985)
Dautmergen Soil		33	62 % YR	T1NO3	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Dautmergen Soil Dautmergen Soil	Radish Radish	31.6 31.2	17 % YR 51 % YR	TINO3	Leaves Tubers	Maturity Maturity	Photometric Photometric	Greenhouse/Soil Pots Greenhouse/Soil Pots	Hoffmann et al. (1982)
		30	37 % YR					and the second s	Hoffmann et al. (1982) Pieper and Austenfeld (1985)
Sand Culture	Pea Pea	30		T1 (NO ₃) ₃	Leaf	ND	AAS	Greenhouse	
Sand Culture		29	32 % YR (N.S.)		Leaf	ND	AAS	Greenhouse	Potsch and Austenfeld (1985)
Sand Culture	Pea Lettuce	29	21 % YR (N.S.)		Leaf	ND	AAS	Geeenhouse	Potsch and Austenfeld (1985)
Dautmergen Soil		28 25	44 % YR 73 % YR	TlNO3	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Dautmergen Soil	Field Bean		% Yield Increase	T1NO3	Leaves	Maturity	Photometric	Greenhouse/Soil Pots	Hoffmann et al. (1982)
Sand Culture	rield beatl	25 2.5		Tl(III) EDTA	Stom	ND	AAS	Greenhouse	Diener and Austenfold (1995)
			(N.S.)	II (IIII) EDTA	o cem	NO	nno.	G. CEHHOUSE	Pieper and Austenfeld (1985)

Table 20. Elevated thallium levels in plants, continued.

Medium	Use	Level (ppm DW)	Hazard Response	Exposure Pathway	Receptor	Duration	Method	Study Setting	Reference
Sand Culture Sand Culture Sand Culture Dautmergen Soil Sand Culture Dautmergen Soil Sand Culture Dautmergen Soil Sand Culture Sand Culture Sand Culture Sand Culture Sand Culture Sand Culture	Field Bean Barley Pea Lettuce Field Bean Radish Field Bean Rield Bean Field Bean Field Bean Field Bean Field Bean Field Bean Field Bean	23 20 (11-45) 20 20 19 18.4 16 8.6 10 8 8 7 6 5	2.3 % YR (N.S.) 23 % YR 14 % YR (N.S.) 39 % YR 4.6 % YR (N.S.) % Yield Increase 10 % YR (N.S.)	T1NO3 T1NO3 T1(NO3)3 T1NO3 T1(NO3)3	Stem Shoot Leaf Leaves Stem Tubers Stem Tubers Leaf Leaf Leaf Leaf	ND 5 Leaf Stage ND Maturity ND Maturity ND Maturity ND	AAS XRFL AAS Photometric AAS Photometric AAS AAS AAS AAS AAS	Greenhouse Greenhouse Greenhouse Greenhouse/Soil Pots Greenhouse/Soil Pots Greenhouse/Soil Pots Greenhouse Greenhouse Greenhouse Greenhouse Greenhouse Greenhouse Greenhouse	Potsch and Austenfeld (1985) Davis et al. (1978) Pieper and Austenfeld (1985) Hoffmann et al. (1982) Pieper and Austenfeld (1985) Hoffmann et al. (1982) Potsch and Austenfeld (1985) Hoffmann et al. (1982) Potsch and Austenfeld (1985) Pieper and Austenfeld (1985)

broccoli, kohlrabi and cabbage accumulated higher levels of thallium than most other vegetables.

Data for elevated thallium levels in plants are limited (Table 20). Up to 2.8 ppm thallium has been reported for plants near industrial sites (potash fertilizer works, smelter and bituminous coal plant) (Smith and Carson 1977a). Scholl and Metzger (1981) reported 22.6 ppm in green kale, 8.5 ppm in savory, 3.1 ppm in turnips, broccoli, kohlrabi and cabbage, and 0.5 ppm in radishes, carrots, onions, lettuce, tomatoes, cucumbers and numerous other vegetables; all grown on soil containing 4.5 ppm thallium. Hoffman et al. (1982) noted very high thallium levels in rapeseed plants and kohlrabi without large decreases in yields (Table 20). It is apparent that thallium uptake and toxicity are very species dependent and that the ability of some species to accumulate very high levels could pose a threat to the food chain. Hazard levels for thallium in soils and plants are presented in Section 3.5.

3.0 HAZARD LEVEL DEVELOPMENT FOR COPPER, MERCURY, SELENIUM, SILVER AND THALLIUM IN SOILS AND PLANTS

The selection of a phytotoxic level for a heavy metal in soil is complicated by the variance of the metal toxicity with soil characteristics and plant species. For example, the soil pH affects the availability of all five metals reviewed in this document. The availability of copper, mercury, silver and possibly thallium increases with decreasing pH. The availability of selenium increases with increasing pH. The pH of surface soils in the Helena Valley project area range from 4.7 to 8.2 with a mean of 7.2 (EPA 1986). The pH range of Helena Valley background surface soil sites is from 7.8 to 8.1. Most of the lower pH values found in the project area are confined to areas in or near the City of East Helena (EPA 1986).

The major complicating factor for the establishment of a critical hazard level in plant tissues is the wide variation observed among different plant species in metal uptake and their sensitivity to phytotoxicity. "It is clear that metal availability depends as much on the crop grown as on total and extractable concentrations of metal in soil" (Carlton-Smith and Davis 1983). The apparent critical toxicity of a given heavy metal in a specific tissue of a specific plant species appears to be relatively independent of different metal forms or the absorption process (Davis et al. 1978). Published phytotoxic levels for soils and plants are given in Tables 21 and 22, respectively. Table 23 presents values believed to be relevant to the Helena Valley study. The following sections describe how the values in Table 23 have been derived.

3.1 Copper Hazard Levels

Reported phytotoxic concentrations of total copper in soil range upward from typical background values (Table 2). A phytotoxic level of 100 ppm has been selected for the Helena Valley. All total soil copper concentrations in excess of 100 ppm, reported in the reviewed literature, were phytotoxic with yield reductions ranging from 14 to 28 percent. The 100 ppm

Notes	Aq	Cu	Нд	Se	Tl	Reference
Notes		60	***************************************			Kovalskiy and Andryomova (1968)
		100	5	10		El-Bassam and Tietjen (1977)
	2	100	0.3	5		Linzon (1978)
		100	2	10	1 9.	Kloke (1979)
		125				Kitagishi and Yamane (1981)
26% Yield Reduction		200				Wallace et al. (1977a)
"Tolerable Margin"					1	Hoffman et al. (1982)
					4.5	Scholl and Metzger (1981)
Solution Culture Alfalfa		100 (mg/L)				Porter and Sheridan (1981)
Maximum Soil Limit for Sludge Application Recommended		50	2			Commission of the European Communities (1982)
Maximum Soil Limit for Sludge Application Mandatory		100				Commission of the European Communities (1982)
Maximum Permissible Levels in Sludges for use on Agricultural Lands	L	500-3000	5-25	14-100	ş ^a	Environmental Protection Services (1984)

Table 22. Plant tissue levels considered to be phytotoxic (ppm dry weight).

r			N.		22	
Notes	Ag	Cu	нд	Se	Tl	Reference
5 Leaf Barley (Range)	4-5	18-21	2-5	7-90	11-45	Davis et al. (1978)
5 Leaf Barley (Mean)	4	20	3	30	20	Davis et al. (1978)
Maize Seedlings			6			Lipsey (1975)
		>20		50-100		Allaway (1968a)
		20				Reuther and Labanauskas (1966)
		20-30				Jones (1972)
		30				Leeper (1972)
5 Leaf Barley		20				Beckett and Davis (1977)
Oats (leaf)		28.8				Wallace et al. (1977a)
	5-10	20-100	1-3	5-30	20	Kabata-Pendias and Pendias (1984)
Rice Grain			0.5			Ishizuka and Tanaka (1962)
Bermudagrass (Fine Sand)-Toxic			6.4			Weaver et al. (1984)
		20				Ratsch (1974)
Bush Bean (stems)	5.1					Wallace et al. (1977d)
Bush Bean (leaves	5.8	29				Wallace et al. (1977d)
		20-40	9			Chaney et al. (1978)
Plantain Herbage/Clover Shoots		10-38				Dijkshoorn et al. (1979)
Rice Leaves		17-26				Chino (1981)
Orange Leaves		>23				Reuther et al. (1958)
Lemon Leaves		>20.0				Haas and Quayle (1935)
Oats (very chlorotic leaves)		37				Hunter and Vergnano (1953)
Snapbean Leaves		20-30				Walsh et al. (1972)
Peach Leaves (indicated as high range)		20-30				Kenworthy (1950)

Table 23. Proposed hazard levels for soils and plants in the Helena Valley study area.

Medium	Diagnostic Level	Site Location	Metal ppm DW							
	Leve1		Copper,	Mercury	Selenium	Silver	Thallium			
otal Soil otal Soil otal Soil otal Soil	Background Background Background Tolerable ^A	US ^C Helena Valley ^D This Report	24 16.3 1-300 50	0.09 0.08 0.005-1.97 2	0.3 0.07 0.005-5.1 ND	0.70 0.20 0.01-5 ND	0.02-2 0.09 0.1-3.0 1			
otal Soil	Phytotoxic ^B		100	5	10	2	10			
otal Plant otal Plant otal Plant otal Plant	Background Background Background Tolerable	Global ^C Helena Valley ^E This Report	1-20 2.0 10	0.03-0.09 0.08 0.001-0.237 0.2	NR NR Ø.ØØ1-84 ND	NR Ø.4 Ø.06-1.4 2	NR NR ND			
otal Plant	Phytotoxic		20	3	400	5	20			

A. Tolerable refers to a soil or plant tissue element concentration that is greater than background, but scientific literature indicates this level has no adverse effect on plant biology.

B. Phytotoxic refers to a soil or plant tissue element concentration that will inhibit plant growth.

C. Kabata-Pendias and Pendias (1984).

D. Surface soil (0-4"), geometric mean, N=3 (EPA 1986).

E. Above ground biomass, average for alfalfa, cereal grains and grasses (EPA 1986).

phytotoxic total soil copper level has been suggested by several authors, including El-Bassan and Tietjen (1977), Linzon (1978), Kabata-Pendias (1979) and Kloke (1979). Baker (1974) reported phytotoxicity when soil levels exceed 150 to 400 ppm total copper and Kitagishi and Yamane (1981) have noted toxicity at soil levels of 125 ppm copper. The 100 ppm total soil copper concentration is the level at which McGrath et al. (1982) noted initial yield reductions in Lolium perenne (perennial ryegrass) when CuSO₄ was added to soil. No data have been found in the reviewed literature for phytotoxic total soil copper levels for alfalfa. Copper tolerant species may not be affected at the 100 ppm total soil copper level.

A tolerable level of 50 ppm total soil copper has been selected based on the reports of no yield loss to occasional small yield reductions noted below this level. This concentration is near the upper end of background levels found for many areas but below the 75 ppm concentration at which McGrath et al. (1982) noted decreased yields of Lolium perenne.

Total soil copper levels in the Helena Valley project area range from 10 to 41 ppm (EPA 1986). The geometric means for total soil copper in the project area and in the background site are 18.3 and 15.0 ppm respectively. Total soil copper levels present in the Helena Valley would not appear to be phytotoxic to crops.

Phytotoxic copper levels in plant tissues have been reported by numerous authors with good agreement. Phytotoxic values for leaves and shoots range from 15 ppm for plantain to 38 ppm for clover (Table 4). Similar values for barley shoots (Davis et al. 1978), rice leaves (Chino 1981), grass shoots (Dijkshoorn et al. 1979), oat leaves (Hunter and Vergnano 1953), and snap beans (Walsh et al. 1972) were 18 to 21 ppm, 17 to 26 ppm, 19 ppm, 37 ppm, and 20 to 30 ppm respectively. Copper levels found in clover were consistently higher than in grasses and copper was tolerated at higher tissue concentrations by clover as compared to grass (Kubota 1983, Dijkshoorn et al. 1979). Background legume tissue concentrations for various

species were red clover (10.0 ppm) > alfalfa (8.8 ppm) > alsike clover (8.3 ppm) > sweetclover (7.9 ppm) = ladino clover (7.9 ppm) > lotus (7.4 ppm) (Kubota 1983). This author reported background copper values in grasses range from 5.9 ppm (smooth brome) to 4.0 ppm (wheat grass). Erdman et al. (1976) found copper levels consistently lower in grasses as opposed to corn and soybeans in several Missouri soils. These data suggest grasses in general will have lower tissue concentrations for a given soil copper level and apparently a lower phytotoxic tissue level.

A plant phytotoxic copper concentration of 20 ppm in leaf or shoot tissue would appear appropriate for the Helena Valley. This concentration may not produce phytotoxicity in alfalfa or other legume crops but is the level at which phytotoxicity may be expected to occur in most cereal crops, many grasses and some vegetables. A potentially useful tool for such an evaluation may be a system developed by Carlton-Smith and Davis (1983). This system presents ordered rankings (league tables) to compare the relative sensitivity of numerous crops to copper toxicity.

A determination of an overall tolerable level in plant tissues is difficult due to apparent differences in the sensitivity of various plant species. The problem is well exemplified by red clover and plantain. The 10.0 ppm background level for red clover (Kubota 1983) is the same level reported to result in a 50 percent yield reduction in plantain herbage (Dijkshoorn et al. 1979). The intermediate range (that level midway between copper deficiency and copper toxicity) values for a large number of fruits and crops commonly exceed 10 ppm with reported values for wheat and oat grain up to 16.7 ppm and 12.1 ppm copper respectively (Reuther and Labanauskas 1966). The level of 10 ppm suggested for East Helena will approximate a tolerable level for cereal grains. A tolerable level in a particular plant species may also be derived through use of a league table system.

3.2 Mercury Hazard Levels

The selection of a hazard level for mercury in soil can not be made with confidence with available data. Any hazard level for mercury should be specific for soil characteristics, mercury compound and plant species. This problem was demonstrated with the work of Weaver et al. (1984). These authors found the phytotoxic total mercury soil level varied from 8 to >50 ppm for bermuda grass, dependent upon the type of soil, with pH values (in the range of 4.7 to 7.7) apparently being insignificant. Levels considered to be phytotoxically excessive have been reported by several review publications (Table 21) and range from 0.3 to 5 ppm. The Environmental Protection Service (1984) gave a range of 5 to 25 ppm for the maximum total mercury content of sludges applied to agricultural lands.

A very tentative hazard level of 5 ppm total soil mercury is recommended for evaluating the Helena Valley data. is below that found by Weaver et al. (1984) to produce reduced plant growth in bermuda grass under their worst case condition. It is probable that levels considerably higher may be appropriate for soils high in clay or organic matter. Of the 160 surface soil samples analyzed from the Helena Valley, 5 samples exceeded 5.0 ppm total soil mercury (EPA 1986). All of these sites were within 0.81 km (0.5 mi) of the East Helena smelter complex. Total mercury levels for surface soil samples at Helena Valley background sites were within the range of typical background levels (Section 2.2.1). A tentative tolerable level of 2 ppm total soil mercury is suggested for the Helena Valley. value is higher than the maximum background value of 0.78 ppm (Table 5). This level is well below the 8 ppm Weaver et al. (1984) found to be toxic to bermuda grass, but the 2 ppm tolerable concentration has little other support.

Phytotoxic hazard level for mercury in plant tissues are better defined than are those for soils. Davis et al. (1978) reported a phytotoxic level of 3 ppm for barley plants in the 5 leaf state using HgCl₂ in a sand culture. Yield reductions of 9.9 and 11 percent resulted in tomato plants with 0.6 to 0.8 ppm

wet weight mercury levels in terminal (newest growth) foliage using methylmercury hydroxide (MMH) (Haney and Lipsey 1973). These authors found the dry matter content of the tomato plants varied between 8.4 and 11.9 percent of the wet weight, with a mean of 10.3 percent. Recalculating MMH concentrations on a dry weight basis indicates the observed yield reductions occurred at tissue mercury concentrations of 5.8 to 7.8 ppm. These values were quite similar to the 8 ppm mercury tissue concentration found to reduce yields of bermuda grass grown in HgCl₂ amended soil (Weaver et al. 1984). These limited data suggest that once absorbed and translocated to the above ground biomass, the phytotoxicity of the various mercury compounds may be similar. Phytotoxic plant tissue concentrations reported in the literature ranged from 0.5 ppm (for rice grain) to 6.4 ppm for bermuda grass foliage (Table 22).

The most appropriate hazard level for mercury in plants in the Helena Valley would appear to be the 3 ppm reported by Davis et al. (1978). This value fits well with the nontoxic mercury level of 2.9 ppm in bermuda grass reported by Weaver et al. (1984) and the 2.3 ppm level found to be nontoxic to alfalfa by Lindberg et al. (1979). A tolerable level of 0.2 ppm mercury in plant tissue is based upon the 0.2 ppm tissue level found to be toxic to bermuda grass under certain conditions (Weaver et al. 1984). Background concentrations near this level have been observed in onions and radishes (Table 7) but this level is 2 to 10 times higher than most observed background levels.

3.3 Selenium Hazard Level

The average background concentration of total soil selenium in the Helena Valley was reported to be 0.07 ppm (EPA 1986). This value is within the expected range of 0.005 to 4.0 ppm for total selenium in soils of the United States (Kabata-Pendias and Pendias 1984). Selenium is not known to retard plant growth at any concentration encountered naturally in soils, but toxicities to certain plants have been produced in a few greenhouse and field plot studies. Kabata-Pendias and Pendias (1984) reported

that total soil selenium levels of 10 and sometimes 5 ppm were phytotoxically excessive. Hurd-Kauer (1934) found that a total soil selenium level of 30 ppm was toxic to wheat seedlings. A growth reduction of buckwheat resulted when total selenium concentrations in the soil ranged from 10.5 to 39.6 ppm (Martin 1936). These buckwheat plants died when soil selenium levels reached 76.6 ppm.

It must be noted that the various forms of selenium available for plant uptake have different degrees of toxicity (Trelease and Beath 1949). This and the limited and conflicting data regarding phytotoxic levels of selenium in soils pose difficulties in proposing hazard level. A tentative value of 10 ppm is suggested as the phytotoxic level for total soil selenium in surface soils of the Helena Valley. No data have been found in the reviewed literature concerning tolerable levels of soil selenium. An estimated value of 5 ppm has been determined intuitively by evaluating the toxic and background levels of selenium in soils of the United States, but no tolerable level for this parameter is recommended because of insufficient data. The total surface soil $(\emptyset-4 \text{ inch})$ selenium value found for the Helena Valley background sites (n=3) is 0.07 ppm (Table 23). Similar values for the entire Helena Valley project area range from 0.07 to 1.30 ppm (EPA 1986).

Total selenium background levels for plant tissue from the United States range from 0.01 to 4.8 ppm (Connor and Shacklette 1975). While there are no reported cases of selenium being toxic to plants growing under natural conditions, there are a few cases of toxicity under experimental conditions. In the review by Kabata-Pendias and Pendias (1984), 5 to 30 ppm in mature leaf tissue was considered phytotoxic. The Environmental Protection Agency (1985) used 191 ppm in tomatoes and 429 ppm selenium in wheat as a toxic level when selenium was added to soil in sewage sludge application. A reduction of buckwheat plant growth occurred when tissue selenium levels ranged from 35 to 124 ppm. Death of these same plants occurred when tissue selenium levels reached 127 ppm (Martin 1936). Very low yields of alfalfa have

occurred when the plant tops contained 360 ppm selenium and 1000 ppm is highly toxic (Soltanpour and Workman 1980). Yopp et al. (1974) reported no injury to wheat that contained 360 ppm total selenium.

The resistance to selenium toxicity ranges so widely among plants that a general toxicity level cannot be estimated with a high degree of confidence. The limited and conflicting data that are available compound this problem. A toxic level of 400 ppm total selenium in plants is recommended for the Helena Valley (Table 23). Only one source has been located that presented evidence of a tolerable level of selenium in vegetation (Yopp et al. 1974). The tolerable level of selenium in vegetation is be estimated at about 300 ppm but no level has been recommended because of insufficient data. Plant tissue selenium concentrations found in the Helena Valley project area range from 0.001 to 84 ppm (Table 23). These concentrations are below most concentrations that have been reported to be phytotoxic (Table 12).

3.4 Silver Hazard Levels

The background range of total surface soil silver in the Helena Valley was reported to be 0.09 to 0.45 ppm with a mean value of 0.20 ppm (EPA 1986). Total soil silver background levels for the entire nation seldom exceed 0.5 ppm (Connor and Shacklette 1975). No first hand research concerning phytotoxic levels of total silver in soils has been found in the reviewed literature. Kabata-Pendias and Pendias (1984) reported that 2.0 ppm total silver in soils was phytotoxically excessive. A tentative value of 2.0 ppm has been selected as the phytotoxic level for total soil silver in the Helena Valley based on this very limited information (Table 23). A tolerable concentration for total soil silver is likely about 1.0 ppm, but this value has little support from the reviewed literature. Total surface soil silver concentrations found for the Helena Valley project area ranged from 0.09 to 46 ppm (EPA 1986).

Background silver concentrations in plant tissue generally range from 0 to 1.0 ppm with most concentrations below the 0.25

ppm level (Table 15). Background silver concentrations in vegetation reported for the Helena Valley ranged from 0.35 to 1.0 ppm (EPA 1986). Data pertinent to the toxicity of silver in plants are also extremely limited. The review by Kabata-Pendias and Pendias (1984) indicated 5 to 10 ppm silver in plant tissue was excessive or toxic. The yield of bush beans was greatly decreased at stem and leaf silver concentrations of 5.1 and 5.8 ppm respectively (Wallace et al. 1977d). No effect in bush bean yield has been noted with stem and leaf tissue levels of 0.8 and 1.0 ppm silver, respectively. Davis et al. (1978) reported that a 10% yield reduction occurred in spring barley with 4.0 ppm silver in the plant tops. With this limited data, a tentative value of 5.0 ppm silver in plant tissue is suggested as the phytotoxic level (Table 23). A tolerable plant tissue silver concentration of 2 ppm is suggested for the Helena Valley based on background levels and limited experimental data.

3.5 Thallium Hazard Levels

Background total soil thallium levels in North America are generally less than 0.5 ppm (Table 17), and typical background total soil thallium concentrations range from 0.02 to 2 ppm (Kabata-Pendias and Pendias 1984). The background surface soil concentration reported for the Helena Valley was 0.09 (EPA 1986). Thallium levels at which phytotoxic symptoms have been noted range from 1 umol/1 (.2 ppm) for corn and sunflowers in solution culture to 1.4 ppm in soil noted by McCool (1933) for damaged wheat plants. McCool (1933) reported wheat plants were killed at a soil thallium level of 28 ppm. Carson and Smith (1977) state "many crop plants are injured by concentrations of about 7 ppm in the soil," and noted toxic effects to tobacco plants at 1 ppm thallium in soil and 0.4 ppm thallium in water. Cataldo and Wildung (1978) found 40 percent of 2.5 ug/l thallium applied to soil was still in soluble form after 13 days. Similar values for arsenic, cadmium, lead and zinc were 8.8, 34, <1, and 8.2 percent respectively. This study suggests that thallium may be proportionately more available to plants than most soil metals. It is

difficult to determine a specific hazard level for thallium in soil due to the wide variation in tolerance and uptake exhibited by various species of plants and due to the scarceness of data. Hoffman et al. (1982) experienced mixed results with total soil thallium levels from 13 to 503 ppm (Table 18). Scholl and Metzger (1981) noted specific toxicity symptoms in some crops grown on a polluted soil containing 4.5 ppm. Total surface soil thallium values reported for the Helena Valley project area ranged from 0.09 to 2.40 ppm (EPA 1986).

A phytotoxic level for total soil thallium of 10 ppm is suggested for the Helena Valley, but has only marginal support from the reviewed literature (Table 23). Scholl and Metzger reported some toxicity symptoms at total soil levels of 4.5 ppm thallium and Hoffman et al. (1982) reported a 23 percent reduction in the yield of lettuce at 13 ppm total soil thallium. The 10 ppm hazard level should be considered very tentative until research provides more information. Hoffman et al. (1982) suggested 1.0 ppm total soil thallium as a "tolerable margin" and, in the absence of contradicting data, this concentration is suggested as the tolerable level for the Helena Valley.

Hydroponic culture experiments with peas and faba beans utilizing thallium+1 and thallium+3 (as TlNO3, Tl+1 EDTA, $T1(NO_3)_3$ and $T1^{+3}$ EDTA) suggest significant yield decreases in peas will occur at leaf thallium levels near 30 ppm for Tl(NO3)3 (Pieper and Austenfeld 1985, Potsch and Austenfeld 1985). studies indicate TlNO3 is less toxic in plant tissue at comparable concentrations than Tl(NO3)3. Pea leaf levels of 75 ppm thallium as TlNO3 were required to produce similar yield reductions experienced with 30 ppm thallium leaf levels using T1(NO₃)₃. Faba beans were apparently highly resistant to thallium toxicity up to the 2.04 ppm used in the hydroponic The maximum faba bean leaf thallium content (8 ppm), using 2.04 ppm thallium as TlNO3 in the hydroponic solution, did not produce significant yield reductions (Potsch and Austenfeld Thallium concentrations up to 27 ppm have been observed in some of the 35 garden species grown in thallium contaminated

soil (Scholl and Metzger 1981). These authors have indicated that some thallium specific symptoms occurred in some species, but no decrease in yields were apparent. Carlson et al. (1975) found a 50 percent reduction in photosynthesis in corn and sunflowers at leaf concentrations of 82 ppm thallium and Bazzaz et al. (1974) have noted a 50 percent reduction in sunflower leaf photosynthesis at a tissue concentration of 63 ppm. Davis et al. (1978) found 11 to 45 ppm thallium in the leaves of 5 leaf stage barely seedling to be toxic and have reported 20 ppm in barley leaf tissue as the "upper critical level" associated with a 10 percent yield reduction in this species. Based on the limited data available, the 20 ppm thallium tissue concentration has been selected as the phytotoxic level for the Helena Valley (Table 23).

A tolerable thallium concentration in plant materials has not been recommended but is likely less than the 5 ppm in leaf tissue that Pieper and Austenfeld (1985) found to produce a 39 percent yield reduction in faba beans. More research is needed to properly define a tolerable thallium level for plants especially for crops typical of the Helena Valley.

4.0 REFERENCES CITED

- Agarwala, S.C., S.S. Bisht and C.P. Sharma. 1977. Relative effectiveness of certain heavy metals in producing toxicity and symptoms of iron deficiency in barley. Canadian Journal of Botany. V. 55, pp. 1299-1307.
- Allaway, W.H. 1968a. Agronomic controls over the environmental cycling of trace elements. Advances in Agronomy. V. 20, pp. 235-274.
- Allaway, W.H. 1968b. Control of the environmental levels of selenium. <u>In:</u> Hemphill, C.C. Ed. Trace Substances in Environmental Health, Vol. 2. University of Missouri. Columbia MO.
- Allaway, W.H., D.P. Moore, J.E. Oldfield and O.H. Muth. 1966. Movement of physiological levels of selenium from soils through plants to animals. Journal of Nutrition. V. 88, pp. 411-418.
- Allaway, W.H. and E.E. Cary. 1964. Determination of submicro gram amounts of selenium in biological materials. Analytical Chemistry V. 36, pp. 1359.
- Baker, D.E. 1974. Copper: soil, water, plant relationships. Federal Proceedings. V. 31, pp. 1188-1193.
- Bazzaz, F.A., R.W. Carlson and G.L. Rolfe. 1974. The effect of heavy metals on plants: Part 1. Inhibition of gas exchange in sunflower by Pb, Cd, Ni and Tl. Environmental Pollution. V. 7, pp. 241-246.
- Battelle Columbus Laboratories. 1977. Multimedia levels of mercury. Environmental Protection Agency. Washington, D.C. NTIS PB-273 201. 133 pp.
- Becket, P.H.T. and R.D. Davis. 1977. Upper critical levels of toxic elements in plants. New Phytologist. 79, 95-106.
- Bennett, A.C. 1971. Toxic effects of aqueous ammonia, copper, zinc, lead, boron and manganese on root growth. <u>In:</u> Carson, E.W. Ed. The Plant Root and Its Environment. Charlottesville University Press of Virginia. pp. 669-683.
- Blackwood, T.R., D.R. Tierney and T.M. Briggs. 1979. Status assessment of toxic chemicals: mercury. Environmental Protection Agency, Washington, D.C. EPA 600/2-79-210; NTIS PB80-146384. pp. 23.
- Bohmer, R.G. and P. Pille. 1977. Determination of thallium in rock and soil samples. Talanta. V. 24(8), pp. 521-523.

- Bowen, H.J.M. 1966. Trace Elements in Biochemistry. Academic Press, London and New York. 241 pp.
- Browning, E. 1961. Toxicity of Industrial Metals. Butterworths, London. 325 pp.
- Bujtas, C. and E. Cseh. 1981. Effects of heavy metals and chelating agents on potassium uptake of cereal roots. Plant and Soil. V. 63, pp. 97-100.
- Bull, K.R., R.D. Roberts, M.J. Inship, and G.T. Godman. 1977.

 Mercury concentration in soils, grass, earthworms and small

 mammals near an industrial emission source. Environmental

 Pollution. V. 12, 135. Cited <u>In</u>: Kabata-Pendias and Pendias,

 1984.
- Cannon, H.E., H.T. Shacklette, and H. Bastron. 1968. Metal absorption by equisetum (horsetail). US Dept. of Interior. Geological Survey Bulletin 1278-A. p. All.
- Cappon, C.J. 1984. Content and chemical form of mercury and selenium in soil, sludge, and fertilizer materials. Water, Air and Soil Pollution. V. 22, pp. 95-104.
- Carey, A.E., J.A. Gowen, T.J. Forehand, H. Tai and G.B. Wiersma. 1980. Heavy metal concentrations in soils of five United States Cities, 1972 urban soils monitoring program. Pesticides Monitoring Journal. V. 13(4), pp. 150-154.
- Carlson, R.W., F.A. Bazzaz, and G.L. Rolfe. 1975. The effect of heavy metals on plants. II. Net photosynthesis and transpiration of whole corn and sunflower plants treated with Pb, Cd, Ni, and Tl. Environmental Research. V. 10. pp. 113-120.
- Carlton Smith, C.H. and R.D. Davis. 1983. Comparative uptake of heavy metals by forage crops grown on sludge-treated soil. IN: Heavy Metals in the Environment. V. 1. CEP Consultants, Edinburgh U.K. pp. 393-396.
- Carson, B.L. and I.C. Smith. 1977. Thallium An appraisal of environmental exposure: Kansas City, Mo. Midwest Research Institute Technical Report 5. (Prepared for National Institute for Environmental Health Science, Research Triangle Park, North Carolina).
- Carter, D.L., C.W. Robbins and M.J. Brown. 1970. Selenium concentrations in forage on some high northwestern ranges. Journal of Range Management. V. 23(4), pp. 234-238.
- CAST. 1976. Application of sewage sludge to cropland:
 Appraisal of potential hazards of the heavy metals to plants
 and animals. Council for Agricultural Science and
 Technology. No. 64. Ames, Iowa. 63 pp.

- Cataldo, D.A. and R.E. Wildung. 1978. Soil and plant factors influencing the accumulation of heavy metals by plants. Environmental Health Perspectives. V. 27, p. 149-159.
- Chaney, R.L. 1984. Potential toxicity of plants and food chain resulting from land treatment of hazardous wastes. Proceedings: Conferences on risk and decision analysis for hazardous waste disposal. Hazardous Waste Control Research Institute. Silver Springs, MD.
- Chaney, R.L., P.T. Hundemann, W.T. Palmer, R.J. Small, M.C. White and A.M. Decker. 1978. Plant accumulation of heavy metals and phytotoxicity resulting from utilization of sewage sludge and sludge composts on cropland. In: Proceedings National Conference Composting Municipal Residues and Sludges. Information Transfer Inc. Rockville, MD. pp. 86-97.
- Chaney, R.L. 1973. Crop and food chain effects of toxic elements in sludges and effluents. <u>In</u>: Proceedings Joint Conference on Recycling Municipal Sludges and Effluents on Land.

 National Association of State University and Land Grant Colleges. Washington, D.C. pp. 129-141.
- Chapmann, H.D., G.F. Liebig, Jr. and A.P. Vanselow. 1940. Some nutritional relationships, as revealed by a study of mineral deficiency and excess symptoms on citrus. Soil Science Society of America Proceedings. V. 4, pp. 196-200.
- Chattopadhyay, A. and R.E. Jervis. 1974. Multielement determination in market-garden soils by instrumental photon activiation analysis. Analytical Chemistry. V. 46(12), pp. 1630-1639.
- Chino, M. 1981. Metal stress in rice plants. <u>In</u>: Kitagishi, K. and I. Yamane Eds. Heavy Metal Pollution in Soils of Japan. Japan Scientific Societies Press. Tokyo. pp. 65-80.
- Commission of the European Communities. 1982. Proposal for a Council Directive on the use of sewage sludge in agriculture. COM(82) 527. Brussels.
- Connor, J.J., J.R. Keith, and B.M. Anderson. 1976. Trace-metal variation in soils and sagebrush in the Powder River Basin, Wyoming and Montana. Journal Research U.S. Geological Survey. V. 4(1), January-February, pp. 49-59.
- Connor, J.J. and H.J. Shacklette. 1975. Background geochemistry of some rocks, soils, plants, and vegetables in the conterminous United States. U.S. Geological Survey Professional Paper 574-F. 168 pp.
- Cook, J. 1977. Environmental pollution by heavy metals. International Journal of Environmental Studies. V. 9, pp. 253-266.

- Cooper, C.F. and W.C. Jolly. 1970. Ecological effects of silver iodide and other weather modification agents: a review.

 Water Resources Research. V. 6(1), pp. 88-98.
- Corey, R.B. 1981. Adsorption vs. precipitation. <u>In</u>: Anderson, M.A. and A.J. Rubin, eds. Adsorption of inorganics at solid-liquid interfaces. Ann Arbor Science Publishers, Inc. Ann Arbor, Michigan. pp. 161-182.
- Cox, D.P. 1979. The distribution of copper in common rocks and ore deposits. In: Nriagu, J.L. Ed. Copper in the Environment. Part 1. Ecological Cycling. John Wiley and Sons Inc. New York. pp. 19-42.
- Crockett, A.B. and R.R. Kinnison. 1979. Mercury residues in soil around a large coal-fired power plant. Environmental Science and Technology V. 13(6), pp. 712-715.
- Cunningham, J.D., J.A. Ryan and D.R. Keeney. 1975a. Phytotoxicity in and metal uptake from soil treated with metal ammended sewage sludge. Journal of Environmental Quality. V. 4 (4). pp. 455-458.
- Cunningham, J.D., D.R. Keeney and J.A. Ryan. 1975b. Phytotoxicity and uptake of metals added to soils as inorganic salts or in sewage sludge. Journal of Environmental Quality. V. 4(4), pp. 460-462.
- Czuba, M. and T.C. Hutchinson. 1980. Copper and lead levels in crops and soils of the Holland Marsh Area Ontario, Journal of Environmental Quality. V. 9 (4), pp. 566-574.
- Davis, R.D., P.H.T. Beckett and E. Wollan. 1978. Critical levels of twenty potentially toxic elements in young spring barley. Plant and Soil. V. 49, pp. 395-408.
- Davis, R.D. and P.H.T. Beckett. 1978. Upper critical levels of toxic elements in plants. II. Critical levels of copper in young barley, wheat, rice, lettuce and ryegrass. New Phytologist V. 80(1), pp. 23-32.
- Dekock, P.C. 1956. Heavy metal toxicity and iron chlorosis. Annuals Botany (London). V. 20, pp. 133-141.
- Dijkshoorn, W., L.W. Van Broekhoven and J.E.M. Lampe. 1979.
 Phytotoxicity of zinc, nickel, cadmium, lead, copper, and chromium in three pasture plant species supplied with graduated amounts from the soil. Netherlands Journal of Agricultural Science. V. 27, pp. 241-253.
- D'Itri, F.M. 1972. The Environmental Mercury Problem. CRC Press. Cleveland, Ohio. 124 pp.

- Dudas, M.J. and S. Pawluk. 1977. Heavy metals in cultivated soils and in cereal crops in Alberta. Canadian Journal of Soil Science V. 57, pp. 329-339.
 - El-Bassam, N. and C. Teitjen. 1977. Municipal sludge as organic fertilizer with special reference to the heavy metals constituents. In: Soil Organic Matter Studies, Vol 2, IAEA, Vienna. 253 pp. Cited In: Kabata-Pendias and Pendias. 1984.
 - Elfving, D.C., W.M. Haschek, R.A. Stehn, C.A. Bache, and D.J. Lisk. 1978. Heavy metal residues in plants cultivated on and in small mammals indigenous to old orchard soils. Archives of Environmental Health. V. 3-4, pp. 95-99.
 - Environmental Protection Agency. 1986. Final draft remedial investigation of soils, vegetation and livestock for ASARCO East Helena Smelter site, East Helena, Montana. Prepared by CH2M Hill, D.J. Dollhopf, D.R. Neuman and R.B. Rennick.
 - Environmental Protection Agency. 1985. Environmental profiles and hazard indices for constituents of municipal sludge: selenium. Office of Water Regulations and Standards, Washington, D.C.
 - Environmental Protection Service. 1984. Manual for land application of treated municipal waste water and sludge. Environment Canada, Environmental Protection Service Report EPS 6-EP-84-1, Ottawa KIA IC8. 216 pp.
 - Erdman, J.A., H.T. Shacklette and J.R. Keith. 1976. Elemental composition of corn, grains, soybean seeds, pasture grasses, and associated soils from selected areas of Missouri. U.S. Geological Survey Professional Paper 954-D. 23 pp.
 - Fleischer, M. 1970. Summary of the literature on the inorganic geochemistry of mercury. <u>In</u>: Mercury in the Environment. U.S. Geological Survey Professional Paper 713. pp. 6-13.
- Fletcher, K. and V.C. Brink. 1969. Content of certain trace elements in range forages from south central British Columbia. Canadian Journal of Plant Science. V. 49, pp. 517-520.
- Floyd, B.F. 1917. Dieback, or exanthema, of citrus trees. Florida University Agricultural Experiment Station Bulletin 140.
- Forbes, R.H. 1917. Certain effects under irrigation of copper compounds upon crops. University of California Publications Agricultural Science. V. 1, pp. 395-494.
- Frank, R., K. Ishida and P. Suda. 1976. Metals in agricultural soils of Ontario. Canadian Journal of Soil Science. V. 56, pp. 181-196.

- Frank, R., K.I Stonefield, and R. Suda. 1979. Metals in agricultural soils of Ontario. II. Canadian Journal of Soil Science. V. 59, pp. 99-103.
- Frear, D.E.H., and L.E. Dills. 1967. Mechanism of the insecticidal action of mercury and mercury salts. Journal of Economic Entomology V. 60(4), pp. 970-974.
- Friberg, L. and J. Vostal. 1972. Mercury in the Environment. CRC Press. Cleveland Ohio. 215 pp.
- Galloway, J.N., J.D. Thornton, S.A. Norton, H.L. Volchok and R.A.N. McLean, 1982. Trace metals in atmospheric depostion: a review and assessment. Atmospheric Environment. V. 16(7), pp. 1677-1700.
- Geilmann W., K. Beyermann, K.H. Neeb and R. Neeb. 1960. Thallium in regelmassig vorhandenes Spurenelement im tierischen und pflanzlichen organismus. Biochem Zeitsch. V. 333, pp. 62-70. In: Chemistry Abstracts V. 55. no 14528f. 1961.
- Gildon, A. and P.B. Tinker. 1983. Interactions of vesiculararbuscular mycorrhizal infection and heavy metals in plants. I. The effects of heavy metals on the development of vesicular-arbuscular mycorrhizas. New Phytologist. V. 95(2), pp. 247-261.
- Gilmour, J.T. and M.S. Miller. 1973. Fate of a mercuricmercurous chloride fungicide added to turfgrass. Journal of Environmental Quality. V. 2(1), pp. 145-148.
- Gladstones, J.S., J.F. Loneragan and W.J. Simmons. 1975.

 Mineral elements in temperate crop and pasture plants. III

 Copper. Australian Journal of Agricultural Research. V.

 26, pp. 113-126.
- Gough, L.P., H.T. Shackette and A.A. Case. 1979. Element concentrations toxic to plants, animals, and man. U.S. Geological Survey Bulletin 1466. U.S. Government Printing Office, Washington D.C. 79 pp.
- Grossenbacher, J.G. 1916. Some bark diseases of citrus trees in Florida. Phytopathology V. 6, pp. 29-50.
- Gupata, U.C. and K.A. Winter. 1975. Selenium content of soils and crops and the effects of lime and sulfur on plant selenium. Canadian Journal Soil Science V. 55, pp. 161-166.
- Haas, A.R.C. and H.J. Quayle. 1935. Copper content of citrus leaves and fruit in relation to exanthema and fumigation injury. Hilgardia V. 9, pp. 143-177.

- Haney, A. and R.L. Lipsey. 1973. Accumulation and effects of methyl mercury hydroxide in terrestrial food chain under laboratory conditions. Environmental Pollution. V 5(4), pp. 305-316.
- Haque, M.A. and V. Subramanian. 1982. Copper, lead and zinc pollution of soil environment. CRC Critical Reviews in Environmental Control. V. 12(1), pp. 13-68.
- Hara, T. and Y. Sonoda. 1979. Comparison of the toxicity of heavy metals to cabbage growth. Plant and Soil. V. 51, pp. 127-133. Cited In: Dijkshoorn et al. 1979.
- Hazlett, D.W., G.K. Rutherford and G.W. VanLoon. 1983. Metal contaminants in surface soils and vegetation as a result of nickel/copper smelting at Coniston, Ontario, Canada. Reclamation and Revegetation Research. V. 2, pp. 123-137.
- Heilman, P.E. and G.T. Ekuan. 1977. Heavy metals in gardens near the Asarco smelter, Tacoma, Washington. EPA-68-01-2989. pp. 86.
- Hitchcock, A.E. and P.W. Zimmerman. 1957. Toxic effects of vapors of mercury and of compounds of mercury on plants. Annals New York Academy of Sciences. V. 65, pp. 474-497.
- Hogan, G.D. and W.E. Rauser. 1979. Tolerance and toxicity of cobalt, copper, nickel, and zinc in clones of Agrostis gigantea. New Phytologist 83. pp. 665-670.
- Hogan G.D., G.M Courtin and W.E. Rauser. 1977. Copper tolerance in clones of <u>Agrostis</u> gigantea from a mine waste site. Canadian Journal of Botany V. 55, pp. 1043-1050.
- Hogg, T.J., J.W.B. Stewart and J.R. Bettany. 1978. Influence of the chemical form of mercury on its adsorption and ability to leach through soils. Journal of Environmental Quality. V. 7(3), pp. 440-444.
- Hoffmann, Von G.G., P. Schweiger and W. Scholl. 1982. Aufnahme von Thallium durch landwirtschaftliche und gartnerische Nutzpflanzen. Landwirt Forschung V. 35(1-2), pp. 45-54.
- Horovitz, C.J., H.H. Schock, and L.A. Horovitz-Kisimova. 1974. The content of scandium, thorium, silver, and other trace elements in different plant species. Plant and Soil. V. 40, pp. 397-403.
- Hunter, J.G. and L. Vergnano. 1953. Trace-element toxicities in oat plants. Annals of Applied Biology. V. 40, pp 761-777.
- Hurd-Karrer, A.M. 1934. Selenium injury to wheat plants and its inhibition by sulphur. Journal of Agricultural Resources V. 49(4), pp. 343-357.

- Hutchinson, T.C. 1979. Copper contamination of ecosystems caused by smelter activities. <u>In: Nriagu, J.O. Ed., Copper in the Environment Part 1: Ecological Cycling.</u> John Wiley and Sons Inc. New York. pp. 451-502.
- Ishida, R.F. and P. Suda. 1976. Metals in agricultural soils of Ontario. Canadian Journal of Soil Science. V. 56, pp. 181-196.
- Ishizuka, Y. and A. Tanaka. 1962. Inorganic nutrition of rice plant (Part 8). Effect of lead, mercury and arsenic levels in culture solution on yields and chemical composition of the plant. Journal Science Soil Manure, Japan. V. 33, pp. 421-423.
- Jarvis, S.C. 1978. Copper uptake and accumulation by perennial ryegrass grown in soil and solution culture. Journal Science Food Agriculture V. 29, p. 12-18.
- Jenkins, D.W. 1980. Biological monitoring of toxic trace
 metals. Environmental Monitoring Systems Laboratory. U.S.
 Environmental Protection Agency. NTIS PB 81-103475 (V. 1),
 PB 81-103491 (V. 2), PB-103509 (V. 3).
- John, M.K. 1972. Mercury uptake from soil by various plant species. Bulletin of Environmental Contamination and Toxicology V. 8(2), pp. 77-80.
- Jones, J.B. 1972. Plant tissue analysis for micronutrients.

 In: Micronutrients in Agriculture. Mortvedt, J.J., D.M.

 Giordano and W.L. Lindsay Eds. Soil Science Society of America. Madison, Wisconsin. 319 pp.
- Kabata-Pendias, A. and H. Pendias. 1984. Trace Elements in Soils and Plants. CRC Press Inc. Boca Raton, Florida. 315 pp.
- Kabata-Pendias, A. 1979. Current problems in chemical degradation of soils. Paper presented at Conference on Soil and Plant Analyses in Environmental Protection, Falenty/Warsaw, October 29. Cited <u>In</u>: Kabata-Pendias and Pendias. 1984.
- Kenworthy, A.L. 1950. Nutrient element composition of leaves from fruit trees. Proceedings American Society of Horticultural Science. V. 55, pp. 41-46.
- Kitagishi, K. and I. Yamane (Eds.) 1981. Heavy Metal Pollution in Soils of Japan. Japan Scientific Societies Press. Tokyo. 302 pp.
- Klein, D.H. and P. Russel. 1973. Heavy metals: Fallout around a power plant. Environmental Science and Technology. V. 7(4), pp. 357-358.

- Klein, D.H. 1972. Mercury and other metals in urban soils. Environmental Science and Technology V. 6(6), pp. 560-562.
- Kloke, A. 1979. Content of arsenic, cadmium, chromium, flourine, lead, mercury and nickel in plants grown on contaminated soil. Paper presented at United Nations ECE Symp. on Effects of Air-Borne Pollution on Vegetation, Warsaw, August 20. 192 pp. Cited In: Kabata-Pendias and Pendias, 1984.
- Kovalskiy, V.V. and G.A. Andryanova. 1968. Trace elements (Cu, Co, Zn, Mo, Mn, B, I, Sr) in soils of USSR. Buryatskoye Knigi Izd. Ulan-ude V. 56. Cited <u>In</u>: Kabata-Pendias and Pendias, 1984
- Kubota, J. 1983. Copper status of United States soils and forage plants. Agronomy Journal. V. 75, pp. 913-918.
- Lagervall, M. and G. Westoo. 1969. G. varFoda. V. 21, pp. 9.

 In: Lofroth, G. 1970. Methylmercury. Swedish Natural
 Science Research Council. Ecological Research Committee
 Bulletin 4, 2nd Edition, pp. 6-10.
- Lakin, H.W. and D.F. Davidson. 1967. The relation of the geochemistry of selenium to its occurrence in soil. <u>In:</u>
 Proceedings Selenium in Biomedicine. Westport, CT pp. 27.
- Lagerwerff, J.V. 1972. Lead, mercury, and cadmium as environmental contaminants. <u>In</u>: Mortvedt, J.J., D.M. Giordano and W.L. Lindsay Eds. Micronutrients in Agriculture. Soil Science Society of America. pp. 593-636.
- Latterell, J.J., R.H. Dowdy and W.E. Larson. 1978. Correlation of extractable metals and metal uptake of snap beans grown on soil amended with sewage sludge. Journal of Environmental Quality. V. 7(3), pp 435-439.
- Laul, J.C., W.C. Weimer and L.A. Rancitelli. 1977. Biogeochemical distribution of rare earths and other trace elements in plants and soils. <u>In:</u> Proceedings of Second Symposium on the Origin of the Elements. UNESCO. Paris, May 10-13.
- Leeper, G.W. 1972. Reactions of heavy metals with soils with special regard to their application in sewage wastes. U.S. Department of the Army, Corps of Engineers. DACW 73-73-C-0026. 70 pp.
- Levesque, M. 1974. Some aspects of selenium relationships in eastern Canadian soils and plants. Canadian Jounal of Soil Science. V. 54, pp. 205-214.

- Lindberg, S.E., D.R. Jackson, J.W. Huckabee, S.A. Jamzen, M.J. Levin and J.R. Lund. 1979. Atmospheric emission and plant uptake of mercury from agricultural soils near the Almaden mercury mine. Journal of Environmental Quality. V. 8(4), pp. 572-578.
- Lindberg, S.E. and R.R. Turner. 1977. Mercury emissions from chlorine-production solid waste deposits. Nature. V. 268(14), pp. 133-136.
- Linzon, S.N. 1978. Phytotoxicology excessive levels for contaminants in soil and vegetation. Report of Ministry of the Environment. Ontario, Canada. Cited <u>In</u>: Kabata-Pendias and Pendias, 1984.
- Lipsey, R.L. 1975. Accumulation and physiological effects of methylmercury hydroxide on maize seedlings. Environmental Pollution. V. 8, pp. 149. Cited <u>In</u>: Kabata-Pendias and Pendias, 1984.
- Logan, P.G., N.W. Lepp and D.A. Phipps. 1983. Thallium uptake by higher plants. <u>In</u>: Heavy Metals in the Environment. Vol. 1. CEP Consultants Ltd. Edinburgh, U.K. pp. 642-645.
- Logan, T.J. and R.L. Chaney. 1983. Utilization of municipal wastewater and sludge on land metals. Proceedings: Utilization of municipal wastewater and sludge on land. University of California, Riverside. pp. 235-323.
- Martin, A.L. 1936. Toxicity of selenium to plants and animals. American Journal Botany V. 23, pp. 471-483.
- McCarthy, J.H. Jr., J.L. Menschke, W.H. Ficklin and R.E. Learned. 1970. Mercury in the atmosphere. <u>In</u>: W.T. Pecora, Director. Mercury in the environment. U.S. Geological Survey Professional Paper 713. pp. 37-39.
- McCarthy, J.H. Jr, W.W. Vaughn, R.E. Learned, and J.L. Menschke. 1969. Mercury in soil, gas and air-a potential tool in mineral exploration. U.S. Geological Survey Circular 609. pp. 16.
- McCool, M.M. 1933. Effect of thallium sulfate on the growth of several plants and on the nitrification in soils. Contributions of the Boyce Thompson Institute. V. 5(3), pp 289-296.
- McGrath, D., R.F. McCormack, G.A. Fleming and D.B.R. Poole. 1982. Effects of applying copper-rich pig slurry to grassland. I. Pot experiments. Irish Journal of Agricultural Research . V. 21(1), pp. 37-48.
- McKeague, J.A. and M.S. Wolynetz. 1980. Background levels of minor elements in some Canadian soils. Geoderma. V. 24, pp. 299-307.

- McKeague, J.A., J.G. Deguardubi and N.S. Witbnetz. 1979. Minor elements in Canadian soils. Agriculture Canada, Ottawa. LRRI Publication 21.
- Miesch, A.T. and C. Huffman, Jr. 1969. Abundance an distribution of Pb, Cd, Zn and As in soils in the vicinity of a smelter in the Helena Valley, MT. Unpublished report. USGS, Denver, CO.
- Mills, J.G. and M.A. Zwarich. 1975. Heavy metal content of agricultural soils in Manitoba. Canadian Journal of Soil Science. V. 55, pp. 295-300.
- Mitchell, R.L. 1971. Trace elements in soils. <u>In</u>: Trace elements in soils and crops. Ministry of Agriculture. Fish Food Technical Bulletin 21. H.M. Stationery Office. London. pp. 8-20. Cited In: Thornton, 1979.
- National Oceanic and Atmospheric Administration. 1983. Climatic Data, Annual Summary, Montana. V. 86(13).
- National Research Council. 1980. Mineral Tolerance of Domestic Animals. National Academy of Sciences, Washington, D.C. 577 pp.
- National Research Council. 1977. Medical and Biologic Effects of Environmental Pollutants: Copper. National Academy of Sciences. Washington, D.C. 115 pp.
- National Research Council. 1976. Medical and Biological Effects of Environmental Pollutants: Selenium. National Academy of Sciences, Washington, D.C.
- Nriagu, J.O. 1979. Copper in the atmosphere and precipitation.

 In: Nriagu, J.O. Ed. Copper in the Environment. Part 1.

 Ecological Cycling. John Wiley and Sons, Inc. New York.

 pp. 43-75.
- Offedal, I. 1940. Untersuchungen uker die Nebenbestandteile von erxmineralien norwegischer zinkblendefuhrender Vorkommen. Vid. Akad. Oslo. Skr. 1(8) pp. 103-105.
- Paasikallio, A. 1981. The effect of soil Ph and Fe on the availability of Se-75 in sphagnum peat soil. Anals Agric. Fenn. V. 20, pp. 15. Cited <u>In</u>: Kabata-Pendias and Pendias 1984.
- Pieper, B. and F.A. Austenfeld. 1985. Phytotoxicity of thallium (T1) in culture solution. Part 2: Effects of T1 (III) on the growth and heavy metal contents of pea and field bean plants. Zeitschrift fur Pflanzenemaehr. Bodenkunke. V. 148 (1). pp. 83-91.

was a way of the

- Pierce, F.J., R.H. Dowdy and D.G. Grigal. 1982. Concentrations of six trace metals in some major Minnesota soil series. Journal Environmental Quality, V. 11(3), pp. 416-422.
- Porter, J.R. and R.P. Sheridan. 1981. Inhibition of nitrogen fixation in alfalfa by arsenate, heavy metals, fluoride, and simulated acid rain. Plant Physiology. V. 68, pp. 143-148.
- Potsch, U. and F.A. Austenfeld. 1985. Phytotoxicity of thallium (T1) in culture solution. Part 1: Effects of T1 (I) on growth and heavy metal contents of pea and field bean plants. Zeitschrift fur Pflanzenemaehr. Bodenkunke. V. 148 (1), pp. 73-82.
- Price, N.O., W.N. Kinkous and R.W. Engel. 1955. Minor element content of forage plants and soils. Journal of Agriculture and Food Chemistry. V. 3, pp. 226-229.
- Ratsch, H.C. 1974. Heavy-metal accumulation in soil and vegetation from smelter emissions. U.S. Environmental Protection Agency. EPA 660/3-74-012. Corvallis, Oregon. pp. 23.
- Reilly, A. and C. Reilly. 1973. Copper-induced chlorosis in Becium homblei (Dewild.) Duvig et Plancke. Plant and Soil. V. 38, pp. 671. Cited <u>In</u>: Kabata-Pendias and Pendias. 1984.
- Reitz, H.J. and N.F. Shimp. 1953. Copper oxide as a soil amendment for citrus. Proceedings Florida State Horticulture Society. V. 66, pp. 37-42.
- Reuther, W. and C.K. Labanauskas. 1966. Copper. <u>In</u>: Chapman, H.D. Ed., Diagnostic Criteria for Plants and Soils. University of California, Riverside, pp. 157-179.
- Reuther, W., T.W. Embleton, and W.W. Jones. 1958. Mineral nutrition of tree crops. Annual Reviews Plant Physiology V. 9, pp. 175-206.
- Reuther, W., D.F. Smith and A.W. Specht. 1952. Accumulation of the major bases and heavy metals in Florida citrus soils in relation to phosphate fertilization. Soil Science. V. 73, pp. 375-381.
- Rosenfeld, I. and O.A. Beath. 1964. Selenium: Geobotany, Biochemistry, Toxicity and Nutrition. Academic Press, New York. 411 pp.
- Scholl, G. and F. Metzger. 1981. Erhebungen uber die Thallium belastung von Nutzpflanzen auf kontaminierten Boden in Raum Lengerich. Landwirtschaftliche Forschung. Special Issue 38. pp. 216-223.

- Shacklette, H.T. and J.G. Boerngen. 1984. Element concentrations in soils and other surficial materials of the conterminous United States. U.S. Geological Survey Professional Paper 1270, 104 pp.
- Shacklette, H.T. 1980. Elements in fruits and vegetables from areas of commercial production in the conterminous United States. U.S. Geological Survey Professional Paper 1178. 138 pp.
- Shacklette, H.T., J.A. Erdman, and T.F Harms. 1978. Trace elements in plant foodstuffs. <u>In</u>: Toxicity of heavy metals in the environments, Part I, Oehme, F.W. Marcel Dekker, New York. Cited <u>In</u>: Kabata-Pendias and Pendias, 1984.
- Shacklette, H.T. 1970. Mercury content of plants: <u>In</u>: Mercury in the Environment. U.S. Geological Survey Professional Paper 713. pp. 35-36.
- Shacklette, H.T. 1965. Element content of bryophytes. U.S. Geological Survey Bulletin 1198-D, 21 pp.
- Singh, B.R. and E. Steinnes. 1976. Uptake of trace elements by barley in zinc-polluted soils: 2. Lead, cadmium, mercury, selenium, arsenic, chromium, and vanadium in barley. Soil Science. V. 121(1), pp. 38-43.
- Smart, N.A. 1968. Use and residues of mercury compounds in agriculture. Residue Reviews. V. 23. pp. 1-36.
- Smith, I.C. and B.L. Carson. 1977a. Trace Metals in the Environment. Vol. 1-Thallium. Ann Arbor Science Publishers, Inc. Ann Arbor, Michigan. 394 pp.
- Smith, I.C. and B.L. Carson. 1977b. Trace Metals in the Environment, Vol 2 - Silver. Ann Arbor Scientific Publications, Ann Arbor, Mich. 469 pp.
- Soltanpour, P.N. and S.M. Workman, 1980. Use of NH₄HCO₃-DTPA soil test to assess availability and toxicity of selenium to alfalfa plants. Communications in Soil Science and Plant Analysis V. 11(12), pp. 1147-1156.
- Sopper, W.E. and E.M. Seaker. 1984. Strip mine reclamation with municipal sludge. Project Summary. EPA-600/52-84-035. 6 pp.
- Stevenson, F.J. and M.S. Ardakani. 1972. Organic matter reactions involving micronutrients in soils. <u>In</u>: Mortvedt, J.J., P.M. Giordano and W.L. Lindsay, eds. Micronutrients in Agriculture. Soil Science Society of America. Madison, Wisc. pp. 79-114.
- Stumm, W. and J.L. Morgan. 1970. Aquatic Chemistry. Wiley Interscience. New York. 583 pp.

- Swaine, D.J. 1955. The trace-element of soils. Commonwealth Bureau Soil Science Technical Communication. No. 48. 157 pp.
- Thornton, I. 1979. Copper in soils and sediments. <u>In</u>: Nriagu, J.O. ed. Copper in the Environment. Part 1. Ecological Cycling. John Wiley and Sons. New York. pp. 171-216.
- Trelease, S.F. and O.A. Beath. 1949. Selenium. Published by the authors, New York. 292 pp.
- Ure, A.M. and J.R. Bacon. 1978. Comprehensive analysis of soils and rocks by spark-source mass spectrometry. Analyst. V. 103, pp. 807-822.
- U.S. Soil Conservation Service. 1981. Average annual precipitation, Montana. U.S.D.A. Soil Conservation Service.

 Bozeman, Montana. Sheet 8.
- Van Ryswyk, A.L., K. Broersm and C.M. Kalnin. 1976. Selenium content of alfalfa grown on orthic grey luvisolic and carbonated orthic gleysolic soils. Canadian Journal of Plant Science V. 56, pp. 753-756.
- Vanselow, A.P. 1965. Silver. <u>In</u>: H.D. Chapman ed. Diagnostic Criteria for Plants and Soils. University of California, Riverside. pp. 405-408.
- Velikii, A.S., V.Y. Volgin and V.V. Ivanov. 1968. Thallium deposits. Chemical Abstracts V. 68. no 4786.
- Vostal, J. 1972. Transport and transformation of mercury in nature and possible routes of exposure. <u>In</u>: Friberg, L. and J. Vostal Eds. Mercury in the Environment. CRC Press. Cleveland, Ohio, pp. 15-27.
- Wallace, A., E.M. Romney, G.V. Alexander and J.E. Kinnear. 1977a. Phytotoxicity and some interactions of essential trace metals iron, manganese, molybdenum, zinc, copper and boron. Communications in Soil Science and Plant Analysis. V. 8(9), pp. 741-750.
- Wallace, A., G.V. Alexander, and F.M. Chaudhry. 1977b. Phytotoxicity of cobalt, vanadium, titanium, silver, and chromium. Communication in Soil Science and Plant Analysis. V. 8(9), pp. 751-756.
- Wallace, A., and E.M. Romney. 1977c. Synergistic trace metal effects in plants. Communications in Soil Science and Plant Analysis. V. 8(9), pp. 699-707.

- Wallace, A., E.M. Romney and J.E. Kinnear. 1977d. Metal interactions in bush bean plants grown in a glasshouse in amended serpentine soils from California. Communication in Soil Science and Plant Analysis. V. 8(9), pp. 727-732.
- Wallace, A., J.W. Cha, F.M. Chaudhry, J. Kinnear, and E.M. Romney. 1977e. Tolerance of rice plants to trace metals. Communications in Soil Science and Plant Analysis. V. 8(9), pp. 809-817.
- Walsh, L.M., W.H. Erhardt and H.D. Seibel. 1972. Copper toxicity in snapbeans (Phaseolus vulgaris L.). Journal of Environmental Quality. V. 1(2), pp. 197-200.
- Ward, N.J., R.R. Brooks and E. Roberts. 1977. Silver in soils, stream sediments, waters and vegetation near a silver mine and treatment plant at Maratoto, New Zealand. Environmental Pollution. V 13, pp. 269-280.
- Warren, H.V., R.E. Delavault, and J. Barakso. 1966. Some observations on the geochemistry of mercury as applied to prospecting. Economic Geology. V. 61, pp. 1010-1028.
- Warren, H.V. and R.E. Delavault. 1950. Gold and silver content of some trees and horsetails in British Columbia. Bulletin Geological Society America. V. 61, pp. 123-128.
- Weaver, R.W., J.R. Melton, D. Wang and R.L. Duble. 1984. Uptake of arsenic and mercury from soil by bermuda grass Cynodon dactylon. Environmental Pollution Series A. V. 33(2), pp. 133-142.
- Wedepohl, K.H. ed. 1978. Handbook of Geochemistry. Vol. 11-5. Springer-Verlag, Berlin. pp. 81-A-1 to 81-I-6.
- Wedepohl, K.H. and J. Zemann. 1974. "Copper". <u>In</u>: K.H. Wedepohl, Ed., Handbook of Geochemistry. Springer-Verlag, Berlin. pp. 29-A-1 to 29-O-1. Cited <u>In</u>: Nriagu, J.O. 1979
- Whanger, P.D. 1974. Bioenvironmental Impact of Selenium.
 National Ecological Research Laboratory, Environmental
 Protection Agency, Corvallis, Oregon.
- Wolnik, K.A., F.L. Fricke, S.G. Copar, G.L. Braude, M.W. Meyer, R.D. Satzger and R.W. Kuennen. 1983. Elements in major raw agricultural crops in the United States. 2. Other elements in lettuce, peanuts, potatoes, soybeans, sweet corn, and wheat. Journal Agricultural Food Chemistry. V. 31, pp. 1244-1249.

- Yopp, J.H., W.F. Schmid and R.W. Holst. 1974. Determination of maximum permissible levels of selected chemicals that exert toxic effects on plants of economic importance in Illinois. Illinois Institute for Environmental Quality. IIEQ Doc. No. 74-33. Cited In: U.S. Environmental Protection Agency. 1985. Environmental profiles and hazard indices for constituents of municipal sludge: selenium. EPA, Office of Water Regulations and Standards, Washington, D.C.
- Zimmerley, S.R. 1947. Thallium. Salt Lake City Division, Metallurgical Branch. U.S. Bureau of Mines. Salt Lake City, Utah.



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